

MICHIGAN DEPARTMENT OF ENVIRONMENTAL QUALITY
WATER RESOURCES DIVISION
JUNE 2015

STAFF REPORT

RECONNAISSANCE SAMPLING OF PERFLUORINATED COMPOUNDS
IN MICHIGAN SURFACE WATERS AND FISH
2010-2014

1. BACKGROUND

Perfluorinated compounds (PFCs) are a class of organic chemicals in which all of the carbon atoms in the molecular backbone are fluorinated. PFCs in the environment are anthropogenic, have been synthesized for over 50 years, and are used in numerous industrial processes and consumer products. In addition, PFCs are persistent, some of the compounds bioaccumulate in the environment, and several have proven to be toxic to birds and mammals in laboratory testing (Giesy and Kannan, 2002). The toxicities of many PFCs have not yet been tested.

PFCs are surfactants used in fire-fighting foams, paints, polishes, and lubricants, and, because of their unique ability to repel both oil and water, have been used in a wide variety of industrial processes and consumer products. Of all the PFCs, perfluorooctanoic acid (PFOA) and perfluorooctane sulfonate (PFOS) have been the subject of the most toxicological work and environmental monitoring. Both compounds were an end product of a manufacturing process, but they can also be generated as byproducts when other fluorinated compounds break down. Since neither PFOA nor PFOS are very volatile, the global distribution of the two might be due to the degradation of more volatile precursor compounds (Giesy and Kannan, 2002).

In 2001, staff of the Michigan Department of Environmental Quality (MDEQ), Water Resources Division, sampled surface water from rivers in different parts of the state for the presence of PFOA and PFOS in cooperation with a Michigan State University professor who was developing an analytical method for these two compounds. There was interest in analyzing for PFCs due to evidence of their persistence, bioaccumulation, potential health effects, and global occurrence, indicating that these compounds might be emerging contaminants of concern (Giesy and Kannan, 2001; 2002). Collection of samples was designed to provide baseline data in surface waters throughout the state (Aiello, 2005). The limited toxicological data available at the time did not suggest that the levels of these two PFCs in Michigan surface waters were a concern.

A decade later, staff of the MDEQ, Remediation and Redevelopment Division, requested assistance from the State of Michigan toxicologists to assess the significance of PFC levels found at the former Wurtsmith Air Force Base (WAFB) in Oscoda, Michigan. This request resulted in an environmental sampling plan to determine the scope of PFC contamination throughout the state, with emphasis on the area surrounding the former WAFB. The contamination of the former WAFB with PFCs was due to the use of fire-fighting foams containing these compounds during training exercises and to extinguish a fire caused by a plane crash in the late 1970s. The contaminated groundwater from the site vents to the Au Sable River via Clark's Marsh.

Extremely high levels of PFOS were found in fillet samples of pumpkinseed (*Lepomis gibbosus*) and bluegill (*L. macrochirus*) collected from Clark's Marsh in 2011, prompting the Michigan Department of Health and Human Services (MDHHS) (previously the Michigan Department of Community Health [MDCH]) to issue a "Do Not Eat" fish consumption advisory for Clark's Marsh. Subsequent advice also stated that consumption of fish from the Au Sable River mouth upstream to the first dam should also be avoided. The discovery of the high levels in fish tissue prompted the MDHHS to apply for funding through the United States Environmental Protection Agency (USEPA), Great Lakes Restoration Initiative (GLRI), to support this survey. The current project was undertaken to reassess some of the surface waters that were initially sampled in 2001, determine the extent of PFC contamination throughout the state, and attempt to correlate surface water concentrations of PFCs with those found in fish tissue.

2. SUMMARY

1. A screening survey of levels of PFCs in Michigan surface waters and fish fillets was prompted by the discovery of high levels of contamination at the WAFB in Oscoda, Michigan.
2. Co-located water and fish samples were collected from 11 sites and analyzed for PFCs. In addition, fish samples collected from 8 Great Lakes and inland sites were analyzed for PFCs.
3. The dominant PFC in water varied between sites. Several of the compounds were detected at most of the sites. PFOS was detected in water at 10 of the 13 sites sampled, and PFOA was detected at all but 1 site.
4. Concentrations of PFOS in river water samples were lower at nearly all of the sites sampled in 2013 compared to concentrations in samples collected at the same sites in 2001.
5. PFOS was detected in over 99% of the fish fillets tested, and made up an average of 82% of the total PFC concentration measured.
6. PFOA was detected in fish only from the Saginaw and Thunder Bay Rivers, and low concentrations were measured there.
7. PFC concentrations in fish were not related to fish length or lipid content. Total PFC and PFOS concentrations were generally highest in the top predator fish at a given site.
8. PFOS concentrations in fish fillets are high enough to cause consumption advisories in 67% of the fish populations sampled for this project. In most cases there are already consumption advisories due to other contaminants.
9. PFOS and PFOA concentrations in water may be at sufficiently high concentrations to impact aquatic life in Clark's Marsh. The PFOS concentrations at 2 sites exceeded the Michigan Rule 57 nondrinking and drinking water human health values. The PFOA concentrations in water and fish are not likely to affect human health.

3. METHODS

Following a USEPA-approved quality assurance project plan, fish and ambient water samples were collected for the GLRI-funded survey from the following water bodies in 2013 and 2014: Au Sable, Flint, Kalamazoo, Muskegon, Saginaw, St. Joseph, Thunder Bay, and Tahquamenon Rivers, and Van Etten Lake (Table 1). In addition to those samples collected for the GLRI survey, fillet samples of fish collected between 2010 and 2012 from Lake Erie, Little Bay De Noc (northern Lake Michigan), the St. Marys River, and 4 inland lakes and impoundments were analyzed for PFCs (Figure 1).

In 2013, water samples were collected from 12 sites (Table 1) following standard operating procedures (U.S. Geological Survey, variously dated). Three surface water grab samples were collected from stream sites near the thalweg at 15-30 centimeters depth using a stainless steel Kemmerer, a pole dipper, or by hand dipping the bottle. A near-surface, mid-depth, and near-bottom grab sample was collected at the deepest point in each of the 2 impoundment sites using a stainless steel Kemmerer.

Surface water samples were collected from the upper pond of Clark's Marsh in 2011 by means of a peristaltic pump with Tygon tubing deployed to mid-depth at 3 sampling points.

Fish were collected using standard electrofishing or netting procedures, and were prepared as standard edible portion samples following the MDEQ, Water Resources Division, Procedure WRD-SWAS-004. The goal for the fish collections was to collect 10 specimens of at least 3 species at each sampling site. The primary target species for the GLRI-funded sampling included common carp (*Cyprinus carpio*), rock bass (*Ambloplites rupestris*), and smallmouth bass (*Micropterus dolomieu*). Carp are insectivores/detritivores and are considered to have a trophic status between level 2 and 3. Rock bass are generally considered to be between trophic level 3 and 4; bluegill and/or pumpkinseed, both trophic level 3 species, were substituted for rock bass at 4 sites. Adult smallmouth bass are considered trophic level 4 (top predators). Largemouth bass (*M. salmoides*) have diets similar to smallmouth bass and are also considered to be a trophic level 4 species; 4 largemouth bass were collected and analyzed along with 8 smallmouth bass from the Thunder Bay River.

The 3 target species were collected at 5 of the GLRI project sites (Table 1). Carp were not collected from either the Au Sable River or Van Etten Lake. Fish collections were completed in 2013 for all GLRI project sites except the Tahquamenon River; sampling was conducted there in 2013 and 2014. Rainbow trout (aka steelhead; *Oncorhynchus mykiss*) and walleye (*Sander vitreus*) were collected from the Au Sable River in 2013 and 2014 and were analyzed for PFCs with funding from the Agency for Toxic Substances and Disease Registry. The latter 2 species were assumed to have migrated upstream from Lake Huron.

Water samples collected in 2013 were analyzed by TestAmerica for 19 PFCs (Table 2). Fish fillet samples from Clark's Marsh were analyzed by Axy's Analytical Services (Sidney, British Columbia); the remaining fillet samples were analyzed by the MDHHS, Analytical Chemistry Laboratory (ACL). PFCs in water and fish samples were analyzed by Liquid Chromatography Tandem Mass Spectrometry. The list of PFCs analyzed varied slightly between labs and over time (Table 2).

The TestAmerica method detection limit for water samples varied by compound from 0.13 to 1.4 nanograms per liter (ng/L), with a reporting limit of 2 ng/L for all PFCs assayed. TestAmerica flagged results below the reporting limit but above the method detection limit; the flagged results were used as reported. The Axy's Analytical Services reporting limit in fish tissue was 0.5 micrograms per kilogram ($\mu\text{g}/\text{kg}$). The MDHHS-ACL limit of detection and reporting limit in fish tissue was 0.25 $\mu\text{g}/\text{kg}$.

Fish tissue concentrations of PFCs were reported as wet weight and entered into the MDEQ, Fish Contaminant Monitoring Program, Access database.

Total PFC values were calculated as the sum of detected values for each individual compound. Values qualified as estimates (J) by the laboratories were treated as the reported value. Non-detect values (K) were assigned a value of zero for the purpose of calculating total

PFC and arithmetic means, and were assigned the detection limit when calculating the geometric mean concentration.

Between-site comparisons of PFC concentrations were made using the Kruskal-Wallis nonparametric test with Dunn's test for multiple comparisons. The statistics were run using the Minitab 15 statistical package.

4. RESULTS AND DISCUSSION

4.1 PFCs in Water

Water samples collected in 2011 from Clark's Marsh had the highest concentrations of total PFC (Table 3); concentrations of nearly all PFCs measured there were at least one order of magnitude higher than the next highest value measured in the 2013 statewide survey. The next highest total PFC concentrations in water samples collected in 2013 for the GLRI project were measured at the Flint River sites, while the Tahquamenon, Thunder Bay, and Muskegon Rivers sites had the lowest total concentrations (Table 3).

The relative contribution of each PFC to the total concentration varied noticeably between sites (Figure 2). The greatest number of PFCs (13) was detected in the Flint River, followed by the Saginaw and St. Joseph Rivers (each with 10). Only 2 PFCs (PFBA and PFHxDA) were detected in Tahquamenon River water samples.

One PFC, (PFBA), had quantifiable concentrations in all water samples from every site sampled. PFBA was produced for use in making photographic film, and may also be a breakdown product of other PFCs. The highest geometric mean concentration of PFBA was measured in Clark's Marsh water samples (116 ng/L), although it made up less than 1% of the total PFC concentration. Water samples from the 12 sites sampled in 2013 had geometric mean concentrations of PFBA ranging from 2.83 to 6.38 ng/L. Michigan does not have drinking or surface water criteria for PFBA. The Minnesota Department of Health established a human health risk limit for PFBA of 7 µg/L (7,000 ng/L) in drinking water, substantially higher than concentrations measured in this survey.

PFOS was detected in water samples collected from Clark's Marsh in 2011 and at 9 of the 12 sites sampled in 2013 (Table 3). The geometric mean concentration in the Clark's Marsh samples was 5,099 ng/L. In comparison, PFOS concentrations at the other 9 sites with detectable levels ranged from 1.4 to 50.7 ng/L. The next highest PFOS surface water concentrations after Clark's Marsh were found in samples from the Flint River; those concentrations were noticeably higher than concentrations measured at any of the other river sites. The concentrations of PFOS measured at Clark's Marsh and at both sampling sites on the Flint River exceeded both the Michigan Rule 57 nondrinking and drinking water human health values of 12 ng/L and 11 ng/L, respectively. Only the Clark's Marsh PFOS concentration exceeded the Rule 57 Final Chronic Value (FCV) (140 ng/L). PFOS was below the limit of detection in samples collected from the Thunder Bay, Tahquamenon, or Muskegon Rivers in 2013.

Concentrations of PFOS in river water samples were lower at nearly all of the sites sampled in 2013 compared to concentrations in samples collected at the same sites in 2001 (Figure 3). The exception was the Flint River where the 2013 PFOS concentration was over 3 times higher than measured in 2001.

PFOA was detected in water samples collected from Clark's Marsh, and at all sites sampled in 2013 except the Tahquamenon River (Table 3). The geometric mean concentration in the Clark's Marsh samples was 1,309 ng/L, and ranged from 1.0 to 4.3 ng/L at the sites sampled in 2013 (Table 3). Aside from Clark's Marsh, PFOA concentrations were highest in the Kalamazoo, Flint, and Saginaw Rivers samples, and the concentrations in samples from those 3 water bodies were roughly equivalent. Geometric mean PFOA concentrations in samples from Van Etten Lake and the Au Sable, Thunder Bay, Muskegon, and St. Joseph Rivers were approximately half the concentrations measured at the other sites sampled in 2013. The Clark's Marsh PFOA concentration did not exceed the nondrinking water human health value, but it did exceed the Rule 57 drinking water value (420 ng/L) and the FCV (880 ng/L). However, since Clark's Marsh is not protected as a drinking water source, the drinking water number does not apply. No samples collected in 2013 had PFOA concentrations exceeding Rule 57 water quality values.

Concentrations of PFOA in river water samples were substantially lower at 7 of the 8 sites sampled in 2013 compared to concentrations at the same sites in 2001 (Figure 3). The exception was the Au Sable River where the 2013 concentration was slightly higher than measured in 2001.

Six PFCs were not detected in any of the 39 water samples, and 1 compound (PFOSA), was detected only in the Clark's Marsh samples (Table 3).

The number of PFCs detected and the total PFC concentrations measured were largely related to the relative urbanization of the watershed, with the exception of the areas sampled near the WAFB (Table 3). The Tahquamenon River flows through primarily forested, undeveloped land in the eastern Upper Peninsula. Two PFCs were detected in the Tahquamenon River water samples, with a total geometric mean of 3.4 ng/L. The Thunder Bay River and upper Muskegon River sample sites also represent areas with relatively low urbanization and had similarly low total PFC concentrations. In contrast, the Flint River sampling site represents a largely urbanized and industrialized watershed; samples taken there had the highest number of PFCs overall and the highest total concentration outside of the WAFB area (Table 3).

4.2 PFCs in Fish Fillets

PFCs have been analyzed in a total of 447 fish fillet samples collected between 2010 and 2013. Five compounds (PFBS, PFBA, PFHpA, PFHxA, and PFPeA) were analyzed only in the 21 samples assayed by Axys Analytical Services, and were not detected in any of those samples (Table 4). Two compounds, PFDS and PFTeA, were not assayed in the 2010 and 2011 samples analyzed by the MDHHS-ACL, but were added to the PFC analyte list for the analysis of the 2012 and 2013 samples.

PFOS was detected in nearly all of the fish fillet samples analyzed (Table 4). The highest PFOS concentrations were measured in pumpkinseed and bluegill from Clark's Marsh, with concentrations ranging from 3,170 to 9,580 µg/kg in the upper pond samples. In addition, 1 pumpkinseed collected from the Au Sable River (near the river mouth) had a PFOS concentration of 2,956 µg/kg. PFOS concentrations measured in all other samples from the Au Sable River, Van Etten Lake, and water bodies statewide ranged from at or below the quantitation limit (0.25 µg/kg) to 610 µg/kg. Mean PFOS concentrations in fish and water are presented in Table 5.

PFOS comprised from 44.5% to 99.2% of the total PFC concentration in fish fillets, and averaged 82% across all species and sites sampled in 2013 (Tables 6-16; Figures 4-14). The highest average PFOS concentration outside of the Oscoda WAFB area was measured in samples from the Flint River, where smallmouth bass had a mean concentration of 132.1 µg/kg (Table 8; Figure 6).

PFOS was also the dominant PFC measured in walleye fillets from 8 selected water bodies sampled between 2010 and 2014 (Figure 15).

Four PFCs were detected in 50 to 60% of the fillet samples (Table 4). Those 4 (PFUnA, PFTeA, PFTriA, and PFDA) comprised 3.5%, 3.3%, 3.3%, and 1.6% of the total PFC concentration on average, respectively.

PFOA was detected in 17% of all fish samples, but only in samples from the Saginaw and Thunder Bay Rivers (Tables 4, 12, and 15; Figures 10 and 13). Overall PFOA comprised less than 1% of the total PFC concentration. The maximum average PFOA concentration of 2.6 µg/kg was measured in smallmouth bass from the Saginaw River.

4.2.1 Factors Affecting PFC Concentrations in Fish Tissue

The concentration of many bioaccumulative contaminants in a fish increases as the fish ages, and comparing concentrations between water bodies should take fish age into account. Determining the age of a fish is somewhat labor intensive; since fish length generally serves as a proxy for fish age, it is substituted when comparing contaminant concentrations between different fish populations.

Fish length versus total PFC concentration was evaluated for several species at all locations either graphically or using linear regression. There was no significant relationship between fish length and total PFC or PFOS concentrations.

Many contaminants are strongly associated with lipids (e.g., PCBs, chlorinated pesticides, dioxins). PFCs, however, tend to bind to proteins rather than fatty tissues, and percent lipid does not act as a good predictor of PFC concentration in fish. There was no significant relationship between fish lipid content and total PFC or PFOS concentrations.

The lack of a relationship between PFC concentration and fish length or lipid content is also reported in work by the Minnesota Pollution Control Agency (MPCA, 2010).

4.2.2 Inter-Species Differences

Rates of contaminant uptake and retention tend to vary between fish species, based on food habits, preferred habitat, and inter-species differences in metabolism and physiology. This tendency holds for at least some PFCs assayed in fish tissue for this study.

Average concentrations of total PFCs and PFOS were highest in the top predator (smallmouth bass) at all sites sampled in 2013 with the exception of the St. Joseph and Au Sable Rivers (Figures 16 and 17); rock bass and bluegill/pumpkinseed, respectively, had higher average concentrations at those 2 sampling sites. Carp had the lowest average total PFC and PFOS concentrations compared to the other species at all 8 sites where carp were collected.

4.2.3 Spatial Differences

Smallmouth bass were collected at all 8 sites sampled in 2013, and all samples had quantifiable levels of PFCs. The smallmouth bass collected from the Au Sable River in 2011 and 2012 also had quantifiable levels of PFCs. PFOS was detected in all smallmouth bass samples. The species provides the best results for use in between-site comparisons.

Concentrations of total PFCs and PFOS in smallmouth bass were highest in the Flint River samples. PFOS concentrations there were statistically different than in smallmouth bass from the Saginaw, St. Joseph, Thunder Bay, Tahquamenon, and Muskegon Rivers (Figure 18).

Walleye collected from 8 sites (Table 1; Figure 19) also provide an indicator of the distribution of PFOS in Michigan waters. The median PFOS concentrations were low in walleye from Hoist Basin (Marquette County), Lake St. Helen (Roscommon County), Otter Lake (Houghton County), and the St. Marys River at Munuscong Bay. These sampling sites represent watersheds with relatively low population density and little industrial activity. Walleye from Lake Erie and Little Bay De Noc (northern Lake Michigan) represent watersheds that have been impacted by manufacturing activity and a higher degree of urbanization; PFOS concentrations in those fish were relatively elevated. The highest medians and greatest range of PFOS concentrations in walleye were measured in samples from Van Etten Lake and the Au Sable River, both likely to have been influenced by legacy contamination.

4.2.4 Fish Consumption Guidance Based on PFOS

MDHHS toxicologists have developed fish consumption screening values (FCSV) for 7 contaminants found in fish, including PFOS (MDCH, 2014a). The MDHHS uses those values to issue consumption advice based on estimated concentrations of the contaminants in sport-caught fish. The protocol for determining a recommended safe rate of consumption includes the calculation of the upper 95% confidence limit (95% UCL) on the mean concentration of a specified contaminant in a fish species collected from a given water body. According to the protocol (MDCH, 2014b) a minimum of 5 fish of legal size are needed for a valid calculation. The 95% UCL is then compared to the FCSV to decide the appropriate safe rate of consumption. The MDHHS developed FCSVs corresponding to fish consumption rates varying from 16 meals per month to a “Do Not Eat” recommendation. The MDHHS does not determine the potential health risk for people consuming more than 16 meals of fish per month.

The 95% UCL of the mean PFOS concentration calculated for the 42 different fish populations sampled are presented in Table 17. The corresponding rates of consumption based on the MDHHS screening values for PFOS are presented in Table 18. Note that these consumption rates are based solely on PFOS concentrations, and that the final consumption advice will be provided by the MDHHS in their “Eat Safe Fish” Guides. The “Eat Safe Fish” guidelines are based on all contaminants analyzed in the samples and in some cases are also subject to management determinations.

PFOS concentrations exceeded the 16 meals per month FCSV (9 µg/kg) in 28 of the 42 (67%) fish populations sampled. Pumpkinseed from Clark’s Marsh and smallmouth bass from the Kalamazoo River at New Richmond had PFOS concentrations exceeding the Do Not Eat FCSV (300 µg/kg). The Kalamazoo River downstream of Morrow Dam to the river mouth, which includes the New Richmond reach, is under a “Do Not Eat” advisory for all species due to legacy PCB contamination. It is unlikely that PFOS will be the primary contaminant driving

consumption advice in any of the waters sampled, with the exception of those waters affected by the WAFB contamination.

4.3 PFC Bioaccumulation

Fillet Bioaccumulation Factors (BAFs) were derived for PFOS and PFOA so that they could be compared to fillet BAFs used to derive Rule 57 human health values. Fillet BAFs were not derived for the other PFCs because they were not routinely detected in fish and water samples and insufficient toxicity data were available to derive Rule 57 human health values.

Table 19 shows the fillet BAFs that were calculated for PFOS using the data presented in this report. The magnitude of the BAFs varied based on species, trophic level, and location. The trophic level 3 BAF for all sites (bluegill/pumpkinseed) was 3,535 liters per kilogram (L/kg), whereas, the trophic level 4 BAF for all sites (averaging values for largemouth/smallmouth bass and rock bass) was 8,371 L/kg. The data suggested that the fillet BAFs may be lower when levels of PFOS in the water column are extremely high. This finding is supported by a study which found that green mussels (*Perna viridis*) exposed to 1 ug/L accumulated 1.6 times more PFOS than mussels exposed to 10 ug/L (Liu et al., 2011). The researchers concluded from their study that PFOS exhibited a nonlinear adsorption mechanism.

Of the 13 sites analyzed in this study, only the Saginaw and Thunder Bay Rivers had measurable levels of PFOA in water and fish filets. BAFs of 845 and 781 L/kg were calculated for smallmouth bass and pumpkinseed, respectively, for the Saginaw River and a BAF of 327 L/kg was calculated for smallmouth bass for the Thunder Bay River. It was unclear why the BAFs for these 2 sites were higher than expected.

5. Water Quality Values

Sufficient data were only available to derive water quality values for PFOS and PFOA. It is unknown whether the other PFCs measured in surface water and fish in this study could enhance the toxicity of PFOS or PFOA.

Provided below is the basis for the human health and aquatic life values for PFOS and PFOA. An estimated wildlife value (WV) was derived for PFOS using the methodology provided in Rule 57 (R 323.1057[3]) of the Part 4 Rules, Water Quality Standards, promulgated under Part 31, Water Resources Protection, of the Natural Resources and Environmental Protection Act, 1994 PA 451, as amended. The surface water concentrations measured in this study were compared to these values to assess whether human health, aquatic life, and wildlife could potentially be impacted.

5.1 Human Health

Human Noncancer Values (HNV)-

An HNV is the concentration of a substance that can be present in surface waters without posing a human health concern following lifetime exposure via the ingestion of 15 grams of fish per day and either 0.01 liters (nondrinking water HNV) or 2 liters (drinking water HNV) of water per day. All surface waters of the state are protected for incidental ingestion of water, whereas, the Great Lakes, connecting channels, and surface waters

used as a drinking water source are protected for the consumption of 2 liters of water per day.

The drinking and nondrinking water HNVs for PFOS are 11 and 12 ng/L, respectively (Appendix A). These 2 values are almost identical because the consumption of fish is the predominant route of exposure to PFOS due to its bioaccumulation potential. Water samples collected from Clark's Marsh and the 2 sites sampled along the Flint River exceeded both human health values for PFOS.

The drinking and nondrinking water HNVs for PFOA are 420 and 12,000 ng/L, respectively (Appendix A). Unlike PFOS, these 2 numbers differ significantly because the consumption of fish is not the predominant route of exposure to PFOA. None of the samples collected during this study exceeded either value.

Fillet BAFs-

As mentioned previously, HNVs protect for people consuming 15 grams of fish per day. The calculations assume that 76% and 24% of the 15 grams of fish consumed per day are from trophic levels 3 (forage fish) and 4 (piscivorous fish), respectively. BAFs relating fish fillet concentrations to water concentrations are used to determine the exposure of people to the contaminant via the consumption of fish. Trophic level 3 and trophic level 4 fillet BAFs of 2,329 and 5,047 L/kg, respectively, were derived for PFOS using samples collected from the Oscoda area (Table 20). These values were used to generate the PFOS HNVs.

The fillet BAFs used to derive the human health values for PFOS were supported by the findings of this study. The additional data were not used to recalculate the fillet BAFs because more samples were available for the Oscoda area and it was more likely that the concentrations in the fish and water samples at this site were in equilibrium.

The limited bioaccumulation data suggest that PFOA is not very bioaccumulative. The current BAF of 4 L/kg was based on a 12-day laboratory study using rainbow trout, which estimated a steady-state value using a kinetic model. A study in Minnesota did not detect PFOA in whole fish exposed to concentrations ranging from 18 to 33 ng/L in Lake Calhoun and a Mississippi pool (STS Consultants, 2007).

The additional data collected in this study were not used to recalculate the fillet BAFs for PFOA used to derive the human health values because few sites had measurable levels of PFOA in both fish and water.

5.2 Aquatic Life

An FCV is the surface water concentration of a substance that is not expected to cause injurious or debilitating effects to aquatic organisms following long-term exposure. The FCVs for PFOS and PFOA are 140 and 880 ug/L, respectively. None of the samples exceeded the FCV for PFOA. Only the water sample from Clark's Marsh exceeded the Rule 57 FCV for PFOS.

An Aquatic Maximum Value (AMV) is the surface water concentration of a substance that an aquatic community can be exposed to briefly without resulting in unacceptable effects.

The AMVs for PFOS and PFOA are 1,600 and 15,000 ug/L, respectively. None of the samples exceeded the AMVs for PFOS or PFOA.

5.3 Wildlife

WVs-

WVs are derived for the bioaccumulative chemicals of concern (BCCs) listed in the Part 4 rules. Even though PFOS is not listed in the Water Quality Standards, WV screening values were derived to aid in this assessment because studies have shown that PFOS has the potential to bioaccumulate to high levels in fish. WV screening levels were not derived for PFOA because it does not have the potential to bioaccumulate in fish to levels that would meet the definition of a BCC.

A WV is the concentration of a substance that can be present in surface waters without causing population-level impacts on piscivorous wildlife like herring gulls, kingfishers, bald eagles, mink, and otters. Since the toxicity of PFOS to these wildlife species has not been examined, it was necessary to base the WVs on toxicity studies conducted on more routinely tested animals. The avian WV was based on a toxicity study in bobwhite quail, whereas, the mammalian WV was based on a toxicity study in rats. Safety factors were used to convert the safe doses in these species to doses that would be protective of each of the species of piscivorous wildlife mentioned above. Fish and water consumption rates for each of the wildlife species were then used to determine a surface water concentration that would be protective of each species of piscivorous wildlife. The geometric mean of the 3 avian values was used to derive the avian WV, whereas, the geometric mean of the 2 mammalian values was used to derive the mammalian WV.

Avian and mammalian WVs of 35 and 84 ng/L, respectively, were derived for PFOS (Appendix A). The levels of PFOS in Clark's Marsh and the 2 sites sampled along the Flint River exceeded these 2 values.

Whole Fish BAFs-

As mentioned previously, a WV is derived to protect piscivorous wildlife like kingfishers, herring gulls, bald eagles, mink, and otters. BAFs relating the concentration of the contaminant in whole fish are used to determine the exposure of wildlife to the contaminant via the consumption of fish. Due to the limited dataset in Michigan, contaminant data from 2 sites in Minnesota were also used to derive BAFs for whole fish. A trophic level 3 BAF of 2,367 L/kg was derived using the data for yellow perch, golden shiner, bluegill sunfish, and white bass (the bass were considered TL3 fish due to their small size). Since no data were available for trophic level 4 fish, the TL4 BAF was derived by multiplying the TL3 BAF for whole fish by the ratio of the TL4 fillet BAF to the TL3 fillet BAF. These values were used to generate the WVs for PFOS (Table 21).

6 CONCLUSIONS

Total PFC concentrations in water varied widely across the state, with concentrations highest in rivers flowing through urban and industrialized areas. The former WAFB area has a known legacy contamination issue and is an exception to this. The composition of the PFC totals also varied significantly across sampling sites, indicating varied sources and magnitudes of contamination. Future source investigations should focus on locations where PFOS-based

firefighting foam may have been used in large quantities, and on sources in urban centers. The co-location of fish and water samples successfully identified the presence and variation of PFCs in Michigan waters, and identified locations that are minimally impacted by PFCs.

The concentrations of PFOS and PFOA in the 13 Michigan surface waters tested are unlikely to impact aquatic life, except in Clark's Marsh. Comparison of surface water concentrations to the estimated WV for PFOS suggest that a more thorough assessment may be warranted to determine whether PFOS is impacting wildlife in the Oscoda area and along the Flint River. Limited ecological risk assessment work is currently being conducted in the Oscoda area by the U. S. Geological Survey. The results of this study might provide some insight into potential for wildlife to be impacted along the Flint River.

The surface water samples taken at the 13 locations listed in Table 1 suggest that human health is not being impacted by PFOA, although it is unclear why levels of PFOA were elevated in some fish species in the Saginaw and Thunder Bay Rivers. The concentrations of PFOS in Clark's Marsh provide support for the current fish consumption advisories in the Oscoda area. The surface water concentration of PFOS and the accompanying fish data for the Flint River suggest that the MDHHS should assess whether fish consumption advisories specific to PFOS are needed (it is possible that the current fish consumption advisory for PCBs in the Flint River would also protect for PFOS).

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ACKNOWLEDGEMENTS

Funding for this project was provided, in part, through a USEPA-GLRI grant to the MDHHS. Additional funds were provided through the Clean Michigan Initiative-Clean Water Fund.

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Table 1. Water and fish collection sites in Michigan for analysis of PFCs, 2010-2014.

Sample Site	Site Selection Rationale	Water Sampled	Number of Fish Sampled by Species							
			Carp	Bluegill / Pumpkinseed	Rock bass	Largemouth / Smallmouth bass	Rainbow trout (Steelhead)	Walleye	White sucker	Yellow perch
Au Sable River near mouth	Known source upstream	X		4	8	20			10	
Flint River at M-13	Elevated PFOS water concentration in 2001	X	no fish collection							
Flint River at Montrose (Genesee County)	Alternate fish collection site	X	10		10	10				
Kalamazoo River at New Richmond	Elevated PFOS water concentration in 2001	X	10	10		10				
Kalamazoo River / Lake Allegan	Upstream/downstream comparisons	X	10	10		10				
Kalamazoo River upstream of Lake Allegan	Downstream comparisons	X	no fish collection							
Muskegon River at Rogers Dam Pond	Low PFOS water concentration in 2001	X	10	10		10				
Saginaw River near mouth	Elevated PFOS water concentration in 2001	X	10		10	10				
St. Joseph River near mouth	Elevated PFOS water concentration in 2001	X	10		10	10				
Tahquamenon River (near mouth)	Low PFOS water concentration in 2001	X	6		8	10				6
Thunder Bay River / Lake Besser	Low PFOS water concentration in 2001	X	10		10	12				
Van Etten Lake (Iosco County)	Near known source	X		10	10			10	10	
Clark's Marsh (Iosco County)	Elevated PFOS water concentration in 2011	X		14						
Dead River / Hoist Basin (Marquette County)	Background fish tissue level							10		
Lake Erie (off Monroe)	Background fish tissue level							10		
Lake Huron (Au Sable River)	Popular sport fish (Human health concern)*						10	14		
Lake Michigan (Little Bay De Noc)	Background fish tissue level							10		
Lake St. Helen (Roscommon County)	Background fish tissue level							10		
Otter Lake (Houghton County)	Background fish tissue level							10		
Rogue River / Rockford Impoundment (Kent County)	Possible legacy source					10			10	
St. Marys River (Munuscong Bay)	Background fish tissue level							8		

Shaded samples indicate analyses funded through EPA-GRI grant. * - analyses funded by Agency for Toxic Substances and Disease Registry

Table 2. PFCs as analyzed by media and laboratory.

Compound	Abbreviation	TestAmerica - Water	Axys - Fish	MDHHS-ACL Fish - 2010 & 2011	MDHHS-ACL Fish - 2012 & 2013
Perfluorobutanoic acid	PFBA	X	X		
Perfluorooctanoic acid	PFOA	X	X	X	X
Perfluorooctane sulfonate	PFOS	X	X	X	X
Perfluoropentanoic acid	PFPeA	X	X		
Perfluorohexanoic acid	PFHxA	X	X		
Perfluoroheptanoic acid	PFHpA	X	X		
Perfluorononanoic acid	PFNA	X	X	X	X
Perfluorodecanoic acid	PFDA	X	X	X	X
Perfluoroundecanoic acid	PFUnA	X	X	X	X
Perfluorododecanoic acid	PFDoA	X	X	X	X
Perfluorotridecanoic acid	PFTriA	X		X	X
Perfluorotetradecanoic acid	PFTeA	X			X
Perfluoro-n-hexadecanoic acid	PFHxDA	X			
Perfluoro-n-octadecanoic acid	PFODA	X			
Perfluorobutane sulfonate	PFBS	X	X		
Perfluorohexane sulfonate	PFHxS	X	X	X	X
Perfluoro-1-heptane sulfonate	PFHpS	X			
Perfluorodecane sulfonate	PFDS	X			X
Perfluorooctane sulfonamide	PFOSA	X	X	X	X

Table 3. Geometric mean concentration (ng/L) of PFCs in surface water samples collected in 2011 and 2013 by sample site. Blank indicates compound was not detected.

Compound	Au Sable River	Clark's Marsh	Flint River - M13	Flint River - Montrose	Kalamazoo River - New Richmond	Kalamazoo River - Lake Allegan	Kalamazoo River - u/s Lake Allegan	Muskegon River	Saginaw River	St. Joseph River	Tahquamenon River	Thunder Bay River	Van Etten Lake
Perfluoro-1-heptanesulfonate (PFHpS)		171		1.11									
Perfluorobutane sulfonate (PFBS)		104	2.36	1.43	0.97	1.03	1.15		1.70	1.36			
Perfluorobutanoic acid (PFBA)	2.83	116	6.38	4.75	3.10	2.83	3.20	3.54	5.09	3.64	3.18	3.06	3.59
Perfluorodecane sulfonate (PFDS)													
Perfluorodecanoic acid (PFDA)		2.5	1.03	0.79						0.47			
Perfluorododecanoic acid (PFDoA)													
Perfluoroheptanoic acid (PFHpA)	0.85	173	2.12	3.45	0.91	1.01	1.04	0.84	1.48			0.93	0.84
Perfluorohexane sulfonate (PFHxS)	4.77	3756	2.25	1.87	1.40	1.58	1.78		2.00	1.08			3.20
Perfluorohexanoic acid (PFHxA)	1.60	922	5.70	8.43	2.19	2.00	2.07		4.33	2.02		0.85	1.16
Perfluoro-n-hexadecanoic acid (PFHxDA)	0.47		0.33	0.65	0.51	0.47	0.54	0.35	0.45	0.94	0.17	0.38	0.49
Perfluoro-n-octadecanoic acid (PFODA)													
Perfluorononanoic acid (PFNA)		24.2	2.43	0.80					1.60				
Perfluorooctane sulfonamide (PFOSA)		172											
Perfluorooctane sulfonate (PFOS)	3.23	5099	40.98	50.66	6.82	6.73	7.62		9.53	1.55			1.37
Perfluorooctanoic acid (PFOA)	1.86	1309	4.19	2.29	3.42	3.92	4.26	1.03	3.10	1.82		1.01	1.20
Perfluoropentanoic acid (PFPeA)	1.26	418	4.40	10.32	1.65	1.66	1.70	1.06	3.34	1.45			1.22
Perfluorotetradecanoic acid (PFTeA)	0.21			0.44						0.38		0.23	0.25
Perfluorotridecanoic acid (PFTriA)													
Perfluoroundecanoic acid (PFUnA)													
TOTAL PFCs	17.10	12266	72.19	87.10	21.00	21.32	23.36	7.14	32.62	15.05	3.35	6.75	13.59

Table 4. PFCs analyzed in fish fillet samples collected in 2010-2013.

	N	N > Detection Limit	% Detected
Perfluorobutane sulfonate (PFBS)	21	0	0
Perfluorobutanoic acid (PFBA)	21	0	0
Perfluorodecane sulfonate (PFDS)	366	155	42.3
Perfluorodecanoic acid (PFDA)	441	232	52.6
Perfluorododecanoic acid (PFDoA)	445	184	41.3
Perfluoroheptanoic acid (PFHpA)	21	0	0
Perfluorohexane sulfonate (PFHxS)	447	45	10.1
Perfluorohexanoic acid (PFHxA)	21	0	0
Perfluoronanoic acid (PFNA)	432	50	11.6
Perfluorooctane sulfonate (PFOS)	447	443	99.1
Perfluorooctanesulfonamide (PFOSA)	447	181	40.5
Perfluorooctanoic acid (PFOA)	447	78	17.4
Perfluoropentanoic acid (PFPeA)	21	0	0
Perfluorotetradecanoic acid (PFTeA)	366	183	50
Perfluorotridecanoic acid (PFTriA)	424	248	58.5
Perfluoroundecanoic acid (PFUnA)	445	262	58.9

Table 5. Mean concentrations of PFOS measured in water and fish fillets collected from Clark's Marsh in 2011 and from several sites statewide in 2012 and 2013.

Sample Site	Collection Date	Geometric Mean PFOS in Water (ng/L)	Arithmetic Mean PFOS in Fish Fillets (ug/kg)						
			All Species Combined	Carp	Largemouth / Smallmouth Bass	Bluegill / Pumpkinseed	Rock Bass	Walleye	White Sucker
Au Sable River - Mill St	7/1/2013	3.2	124.4	--	80.1	771	24.7	--	33.9
Clark's Marsh	8/16/2011	5099	5498	--	--	5498	--	--	--
Flint River - M13	8/2/2013	41	--	--	--	--	--	--	--
Flint River - Montrose	10/31/2013	50.7	92.8	51.6	132.1	--	94.8	--	--
Kalamazoo River-New Richmond	6/11/2013	6.8	61.6	28.1	111.2	45.4	--	--	--
Kalamazoo-Lake Allegan	6/11/2013	6.7	52.2	36.4	82.5	37.5	--	--	--
Kalamazoo River u-s Lake Allegan	6/11/2013	7.6	--	--	--	--	--	--	--
Muskegon River - Hersey	6/6/2013	1.3 (ND)	1.8	1.0	1.3	3.1	--	--	--
Saginaw River - Essexville	8/2/2013	9.5	22.8	13.8	31	23.4	--	--	--
St. Joseph River - Riverview Park	9/11/2013	1.6	31.1	21.0	21.7	--	50.7	--	--
Tahquamenon River - mouth	6/18/2013	1.3 (ND)	1.9	0.5	2.2	--	2.5	--	--
Thunder Bay River - Lake Besser	7/1/2013	1.3 (ND)	1.7	0.3	2.7	--	1.8	--	--
Van Etten Lake	7/1/2013	1.4	13.9	--	--	8.4	10.8	23.7	12.6

Table 6. Individual PFCs in water and fish samples collected from the Au Sable River between 2011 and 2013. Concentrations in fish are arithmetic means ($\mu\text{g}/\text{kg}$), water concentrations are geometric means (ng/L). Blank cells indicate concentrations below the detection limit.

Compound	Smallmouth Bass (N=20)	Rock Bass (N=8)	Bluegill/Pumpkinseed (N=4)	Water (N=3)
Perfluorodecane sulfonate (PFDS)		0.11		
Perfluorodecanoic acid (PFDA)	0.93	0.18	0.44	
Perfluorododecanoic acid (PFDoA)	0.59	0.19	0.28	
Perfluorohexane sulfonate (PFHxS)		0.05	0.11	4.77
Perfluorononanoic acid (PFNA)				
Perfluorooctane sulfonate (PFOS)	80.1	24.7	771.2	3.2
Perfluorooctane sulfonamide (PFOSA)	0.06		3.85	
Perfluorooctanoic acid (PFOA)				1.86
Perfluorotetradecanoic acid (PFTeA)	0.42	0.03	0.22	0.21
Perfluorotridecanoic acid (PFTriA)	1.31	1.01	0.88	
Perfluoroundecanoic acid (PFUnA)	2.00	0.60	0.34	

Table 7. Individual PFCs in water and fish samples collected from Clark's Marsh in 2011. Concentrations in fish are arithmetic means ($\mu\text{g}/\text{kg}$), water concentrations are geometric means (ng/L). Blank cells indicate concentrations below the detection limit.

Compound	Pumpkinseed (N=14)	Water (N=3)
Perfluorodecane sulfonate (PFDS)		
Perfluorodecanoic acid (PFDA)	1.62	2.8
Perfluorododecanoic acid (PFDoA)		
Perfluorohexane sulfonate (PFHxS)	24.57	3756
Perfluorononanoic acid (PFNA)	0.30	24.2
Perfluorooctane sulfonate (PFOS)	5498	5099
Perfluorooctane sulfonamide (PFOSA)	91.46	171.7
Perfluorooctanoic acid (PFOA)	0.86	1309
Perfluorotetradecanoic acid (PFTeA)		
Perfluorotridecanoic acid (PFTriA)		
Perfluoroundecanoic acid (PFUnA)	0.93	

Table 8. Individual PFCs in water and fish samples collected from the Flint River in 2013. Concentrations in fish are arithmetic means ($\mu\text{g}/\text{kg}$), water concentrations are geometric means (ng/L). Blank cells indicate concentrations below the detection limit.

Compound	Carp (N=10)	Smallmouth Bass (N=10)	Rock Bass (N=10)	Water (N=3)
Perfluorodecane sulfonate (PFDS)	0.95	0.76	0.89	
Perfluorodecanoic acid (PFDA)	0.80	2.20	1.62	0.79
Perfluorododecanoic acid (PFDoA)	0.93	2.32	1.50	
Perfluorohexane sulfonate (PFHxS)				1.87
Perfluorononanoic acid (PFNA)				0.80
Perfluorooctane sulfonamide (PFOSA)	0.30	0.47	0.36	
Perfluorooctane sulfonate (PFOS)	51.6	132.1	94.8	50.7
Perfluorooctanoic acid (PFOA)				2.29
Perfluorotetradecanoic acid (PFTeA)		0.68	0.60	0.44
Perfluorotridecanoic acid (PFTriA)		0.78	0.58	
Perfluoroundecanoic acid (PFUnA)	0.49	2.32	1.36	

Table 9. Individual PFCs in water and fish samples collected from the Kalamazoo River at Lake Allegan in 2013. Concentrations in fish are arithmetic means ($\mu\text{g}/\text{kg}$), water concentrations are geometric means (ng/L). Blank cells indicate concentrations below the detection limit.

Compound	Carp (N=10)	Smallmouth Bass (N=10)	Bluegill (N=10)	Water (N=3)
Perfluorodecane sulfonate (PFDS)	0.68	0.64	0.22	
Perfluorodecanoic acid (PFDA)	0.58	0.61		
Perfluorododecanoic acid (PFDoA)	0.24	0.04		
Perfluorohexane sulfonate (PFHxS)				1.58
Perfluorononanoic acid (PFNA)				
Perfluorooctane sulfonate (PFOS)	36.4	82.5	37.5	6.7
Perfluorooctane sulfonamide (PFOSA)	0.78	1.63	0.28	
Perfluorooctanoic acid (PFOA)				3.92
Perfluorotetradecanoic acid (PFTeA)		0.04	1.18	
Perfluorotridecanoic acid (PFTriA)				
Perfluoroundecanoic acid (PFUnA)	0.36			

Table 10. Individual PFCs in water and fish samples collected from the Kalamazoo River at New Richmond in 2013. Concentrations in fish are arithmetic means ($\mu\text{g}/\text{kg}$), water concentrations are geometric means (ng/L). Blank cells indicate concentrations below the detection limit.

Compound	Carp (N=10)	Smallmouth Bass (N=10)	Bluegill (N=10)	Water (N=3)
Perfluorodecane sulfonate (PFDS)	0.52	1.69	0.18	
Perfluorodecanoic acid (PFDA)	0.03	2.09	0.17	
Perfluorododecanoic acid (PFDoA)		1.06	0.09	
Perfluorohexane sulfonate (PFHxS)				1.4
Perfluorononanoic acid (PFNA)				
Perfluorooctane sulfonate (PFOS)	28.1	111.2	45.4	6.8
Perfluorooctane sulfonamide (PFOSA)	0.51	2.92	0.52	
Perfluorooctanoic acid (PFOA)				3.42
Perfluorotetradecanoic acid (PFTeA)		0.66	0.82	
Perfluorotridecanoic acid (PFTriA)		1.89	0.43	
Perfluoroundecanoic acid (PFUnA)		2.33	0.08	

Table 11. Individual PFCs in water and fish samples collected from the Muskegon River in 2013. Concentrations in fish are arithmetic means ($\mu\text{g}/\text{kg}$), water concentrations are geometric means (ng/L). Blank cells indicate concentrations below the detection limit.

Compound	Carp (N=10)	Smallmouth Bass (N=10)	Bluegill (N=10)	Water (N=3)
Perfluorodecane sulfonate (PFDS)	0.04			
Perfluorodecanoic acid (PFDA)		0.05	0.04	
Perfluorododecanoic acid (PFDoA)	0.03			
Perfluorohexane sulfonate (PFHxS)				
Perfluorononanoic acid (PFNA)				
Perfluorooctane sulfonate (PFOS)	1.0	1.3	3.1	
Perfluorooctane sulfonamide (PFOSA)				
Perfluorooctanoic acid (PFOA)				1.03
Perfluorotetradecanoic acid (PFTeA)		0.03	0.3	
Perfluorotridecanoic acid (PFTriA)		0.03	0.03	
Perfluoroundecanoic acid (PFUnA)		0.3	0.1	

Table 12. Individual PFCs in water and fish samples collected from the Saginaw River in 2013. Concentrations in fish are arithmetic means ($\mu\text{g}/\text{kg}$), water concentrations are geometric means (ng/L). Blank cells indicate concentrations below the detection limit.

Compound	Carp (N=10)	Smallmouth Bass (N=10)	Pumpkinseed (N=10)	Water (N=3)
Perfluorodecane sulfonate (PFDS)	0.7	0.5		
Perfluorodecanoic acid (PFDA)	0.11	4.53	3.18	
Perfluorododecanoic acid (PFDoA)	0.11	3.95	3.41	
Perfluorohexane sulfonate (PFHxS)				2.0
Perfluorononanoic acid (PFNA)		2.91	2.94	1.60
Perfluorooctane sulfonate (PFOS)	13.8	31.0	23.4	9.5
Perfluorooctane sulfonamide (PFOSA)	0.03	0.42	0.38	
Perfluorooctanoic acid (PFOA)		2.62	2.42	3.10
Perfluorotetradecanoic acid (PFTeA)		5.88	5.91	
Perfluorotridecanoic acid (PFTriA)		8.52	6.17	
Perfluoroundecanoic acid (PFUnA)	0.38	6.93	4.29	

Table 13. Individual PFCs in water and fish samples collected from the St. Joseph River in 2013. Concentrations in fish are arithmetic means ($\mu\text{g}/\text{kg}$), water concentrations are geometric means (ng/L). Blank cells indicate concentrations below the detection limit.

Compound	Carp (N=10)	Smallmouth Bass (N=10)	Rock Bass (N=10)	Water (N=3)
Perfluorodecane sulfonate (PFDS)	0.9	0.43	1.75	
Perfluorodecanoic acid (PFDA)		0.18	0.5	0.47
Perfluorododecanoic acid (PFDoA)	0.58		0.13	
Perfluorohexane sulfonate (PFHxS)				1.08
Perfluorononanoic acid (PFNA)		0.04		
Perfluorooctane sulfonate (PFOS)	21.0	21.7	50.7	1.6
Perfluorooctane sulfonamide (PFOSA)	0.03	0.57	0.51	
Perfluorooctanoic acid (PFOA)				1.82
Perfluorotetradecanoic acid (PFTeA)		0.84	2.13	0.38
Perfluorotridecanoic acid (PFTriA)		0.7	1.61	
Perfluoroundecanoic acid (PFUnA)	0.1	0.24	0.05	

Table 14. Individual PFCs in water and fish samples collected from the Tahquamenon River in 2013. Concentrations in fish are arithmetic means ($\mu\text{g}/\text{kg}$), water concentrations are geometric means (ng/L). Blank cells indicate concentrations below the detection limit.

Compound	Carp (N=6)	Smallmouth Bass (N=10)	Rock Bass (N=8)	Yellow Perch (N=6)	Water (N=3)
Perfluorodecane sulfonate (PFDS)					
Perfluorodecanoic acid (PFDA)		0.36	0.11		
Perfluorododecanoic acid (PFDoA)		0.32	0.03		
Perfluorohexane sulfonate (PFHxS)					
Perfluorononanoic acid (PFNA)			0.03		
Perfluorooctane sulfonate (PFOS)	0.5	2.2	2.5	1.3	
Perfluorooctane sulfonamide (PFOSA)					
Perfluorooctanoic acid (PFOA)					
Perfluorotetradecanoic acid (PFTeA)	0.06	0.47	0.38		
Perfluorotridecanoic acid (PFTriA)	0.05	0.77	0.35		
Perfluoroundecanoic acid (PFUnA)		0.82	0.51		

Table 15. Individual PFCs in water and fish samples collected from the Thunder Bay River in 2013. Concentrations in fish are arithmetic means ($\mu\text{g}/\text{kg}$), water concentrations are geometric means (ng/L). Blank cells indicate concentrations below the detection limit.

Compound	Carp (N=10)	Smallmouth Bass (N=12)	Rock Bass (N=10)	Water (N=3)
Perfluorodecane sulfonate (PFDS)				
Perfluorodecanoic acid (PFDA)		0.23	0.03	
Perfluorododecanoic acid (PFDoA)		0.49		
Perfluorohexane sulfonate (PFHxS)				
Perfluorononanoic acid (PFNA)				
Perfluorooctane sulfonate (PFOS)	0.3	2.7	1.8	
Perfluorooctane sulfonamide (PFOSA)		0.08		
Perfluorooctanoic acid (PFOA)		0.33		1.01
Perfluorotetradecanoic acid (PFTeA)		0.28	0.61	0.23
Perfluorotridecanoic acid (PFTriA)		0.34	0.55	
Perfluoroundecanoic acid (PFUnA)	0.03	0.62		

Table 16. Individual PFCs in fish samples collected from Van Etten Lake in 2010 and 2012, and water samples collected in 2013. Concentrations in fish are arithmetic means ($\mu\text{g}/\text{kg}$), water concentrations are geometric means (ng/L). Blank cells indicate concentrations below the detection limit.

Compound	Pumpkinseed (2012; N=10)	Rock Bass (2012; N=10)	Walleye (2010; N=10)	White Sucker (2010; N=10)	Water (2013; N=3)
Perfluorodecane sulfonate (PFDS)					
Perfluorodecanoic acid (PFDA)		0.08	0.05	0.06	
Perfluorododecanoic acid (PFDoA)					
Perfluorohexane sulfonate (PFHxS)			0.03	0.82	3.20
Perfluorononanoic acid (PFNA)					
Perfluorooctane sulfonate (PFOS)	8.4	10.8	23.7	12.6	1.4
Perfluorooctane sulfonamide (PFOSA)					
Perfluorooctanoic acid (PFOA)				0.34	1.20
Perfluorotetradecanoic acid (PFTeA)	0.22	0.02			0.25
Perfluorotridecanoic acid (PFTriA)	0.42	0.47	0.56	0.55	
Perfluoroundecanoic acid (PFUnA)	0.29	0.27	0.18	0.19	

Table 17. The 95% UCL on the mean PFOS concentration ($\mu\text{g}/\text{kg}$) in fish fillets sampled in 2010-2013.

	Carp	Smallmouth/Largemouth Bass	Rock Bass	Bluegill/Pumpkinseed	Yellow Perch	Steelhead	Walleye	White Sucker
Au Sable River		157	37.2				26	62
Clark's Marsh				5619				
Flint River	64	154	120					
Hoist Basin							1.0	
Kalamazoo River (New Richmond)	38	329		56				
Lake Allegan	43	93		52				
Lake Erie							26	
Lake Huron						19		
Lake St. Helen							2.5	
Little Bay de Noc							12	
Muskegon River	1.6	1.8		4				
Otter Lake							7	
Rogue River		52						26
Saginaw River	18	43		31				
St. Joseph River	25	28	79					
St. Marys River							8	
Tahquamenon River	0.9	2.6	3.3		1.6			
Thunder Bay River	0.4	3.5	2.2					
Van Etten Lake			14	10			33	18
<i>Blank cells indicate no sample taken</i>								

Table 18. Maximum number of meals per month recommended by the MDHHS based solely on PFOS concentrations in fish filets sampled in Michigan waters between 2010 and 2013. (Actual advisories may be based on other chemicals. Check the Eat Safe Fish Guides for the current recommendations).

	Carp	Smallmouth / Largemouth Bass	Rock Bass	Bluegill/Pumpkinseed	Yellow Perch	Steelhead	Walleye	White Sucker
Au Sable River		0.5	4				4	2
Clark's Marsh				DNE				
Flint River	2	0.5	1					
Hoist Basin							16	
Kalamazoo (New Richmond)	2	DNE		2				
Lake Allegan	2	1		2				
Lake Erie							4	
Lake Huron						4		
Lake St. Helen							16	
Little Bay De Noc							12	
Muskegon River	16	16		16				
Otter Lake							16	
Rogue River		2						4
Saginaw River	8	2		4				
St. Joseph River	4	4	1					
St. Marys River							16	
Tahquamenon River	16	16	16		16			
Thunder Bay River	16	16	16					
Van Etten Lake			8	12				8
<i>Blank cells indicate no data for the species in the water body</i>						DNE – Do Not Eat		

Table 19. BAFs for PFOS using water and fish fillet samples collected from Clark's Marsh in 2011 and from other Michigan sites in 2013.

Sampling Site	Collection Date	Geometric Mean PFOS in Water (ng/L)	PFOS Fish Fillet BAFs (L/kg)				
			All Fish Combined	Carp	Largemouth / Smallmouth Bass	Bluegill / Pumpkinseed	Rock Bass
Clark's Marsh	8/16/11	5099	1078	--	--	1078	--
Flint River - Montrose	10/31/13	50.7	1830	1018	2606	--	1870
Flint River - M13	8/2/13	41	--	--	--	--	--
Saginaw River - Essexville	8/2/13	9.5	2400	1453	3263	2463	--
Kalamazoo River u-s Lake Allegan	6/11/13	7.6	--	--	--	--	--
Kalamazoo River-New Richmond	6/11/13	6.8	9059	4132	16353	6676	--
Kalamazoo-Lake Allegan	6/11/13	6.7	7791	5433	12313	5597	--
Au Sable River - Mill St	7/1/13	3.2	38875	--	25031	240938	7719
St. Joseph River - Riverview Park	9/11/13	1.6	19438	13125	13563	--	31688
Van Etten Lake	7/1/13	1.4	9928		--	6000	7714
Thunder Bay River - Lake Besser	7/1/13	< 1.3 (ND)	>1308	>231	>2077	--	>1308
Tahquamenon River - mouth	6/18/13	< 1.3 (ND)	>1462	>385	>1692	--	>1923
Muskegon River - Hersey	6/6/13	< 1.3 (ND)	>1385	>769	>1000	>2385	--
Geometric Mean (> values not used in calculation)			4630*	3372	9135	3535*	7671

* Au Sable River bluegill/pumpkinseed BAF considered an outlier so not used in calculation. ND – below detection limit

Table 20. BAFs for PFOS using water and fish fillet samples collected from the Oscoda area of Michigan from 2011 to 2013.

Sampling Site	Water Sampling Date(s)	Geometric Mean PFOS in Water (ng/L)	PFOS Fish Fillet BAFs (L/kg)			
			Pumpkinseed	Largemouth Bass	Smallmouth Bass	Rock Bass
Upper Clark's Marsh	8/16/11; 11/2/12	3179	1502	1442	--	--
Middle Clark's Marsh	10/29/12; 11/2/12	1887	1883	--	--	--
Lower Clark's Marsh	8/16/11; 10/29/12	99	4465	9141	--	--
Au Sable River	7/1/13	3.2	--	--	7500	6563
Geometric Mean PFOS BAF (L/kg)			2329	3631	7500	6563

Table 21. BAFs for PFOS using water and whole fish samples collected from Clark's Marsh (Michigan), Lake Calhoun (Minnesota), and a Mississippi River Pool (Minnesota).

Sampling Site	Water Sampling Date(s)	Geometric Mean PFOS in Water (ng/L)	PFOS Whole Fish BAFs (L/kg)			
			Yellow Perch	Golden Shiner	Bluegill Sunfish	White Bass
Upper Clark's Marsh	8/16/11; 11/2/12	3179	922; 865	1519	--	--
Middle Clark's Marsh	10/29/12; 11/2/12	1887	938	--	--	--
Lake Calhoun	2006	110	--	--	4113	--
Mississippi River Pool	2006	25.5	--	--	--	5529
Geometric Mean PFOS BAF (L/kg)			908	1519	4113	5529

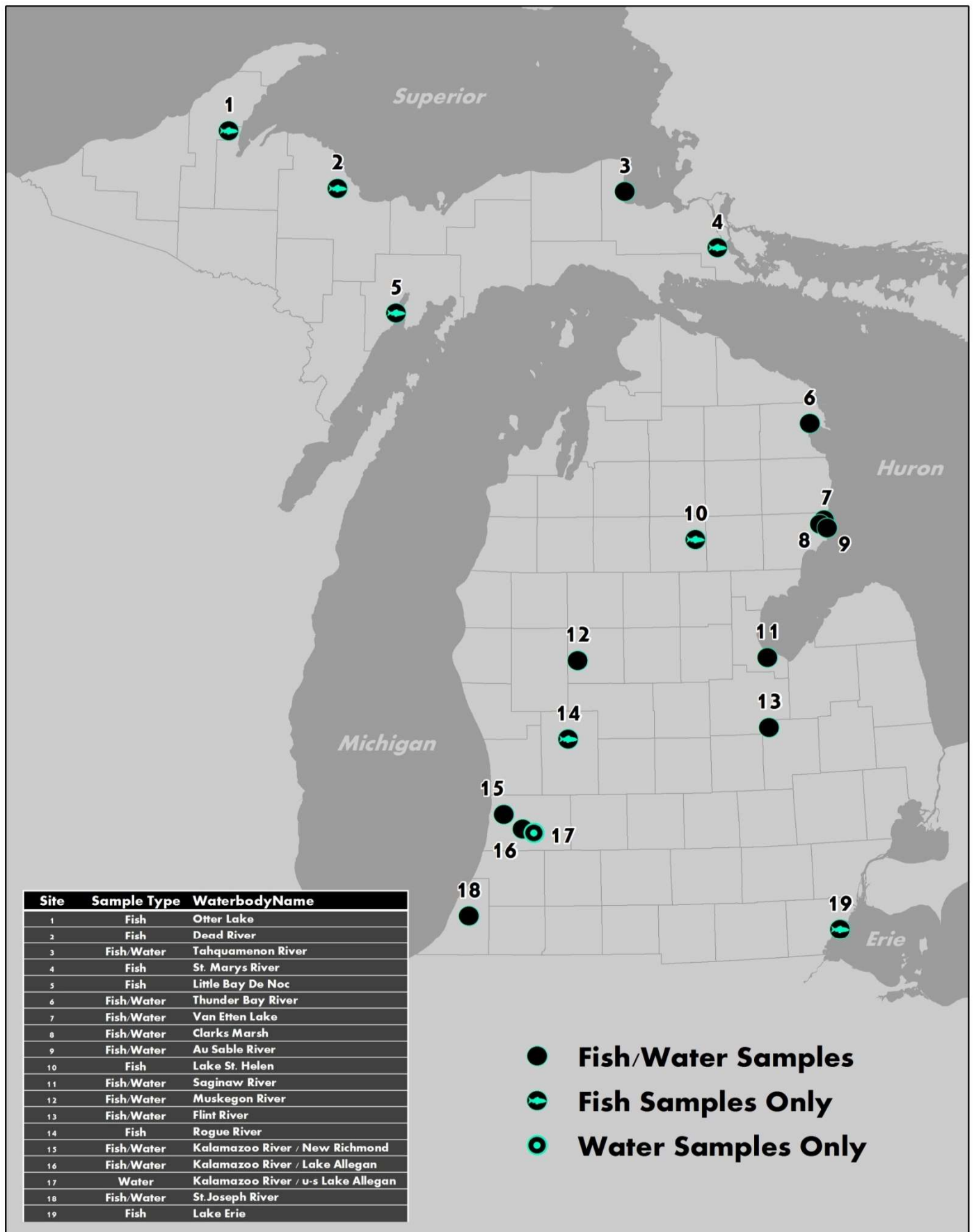


Figure 1. PFC water and fish tissue sample locations, 2010-2014.

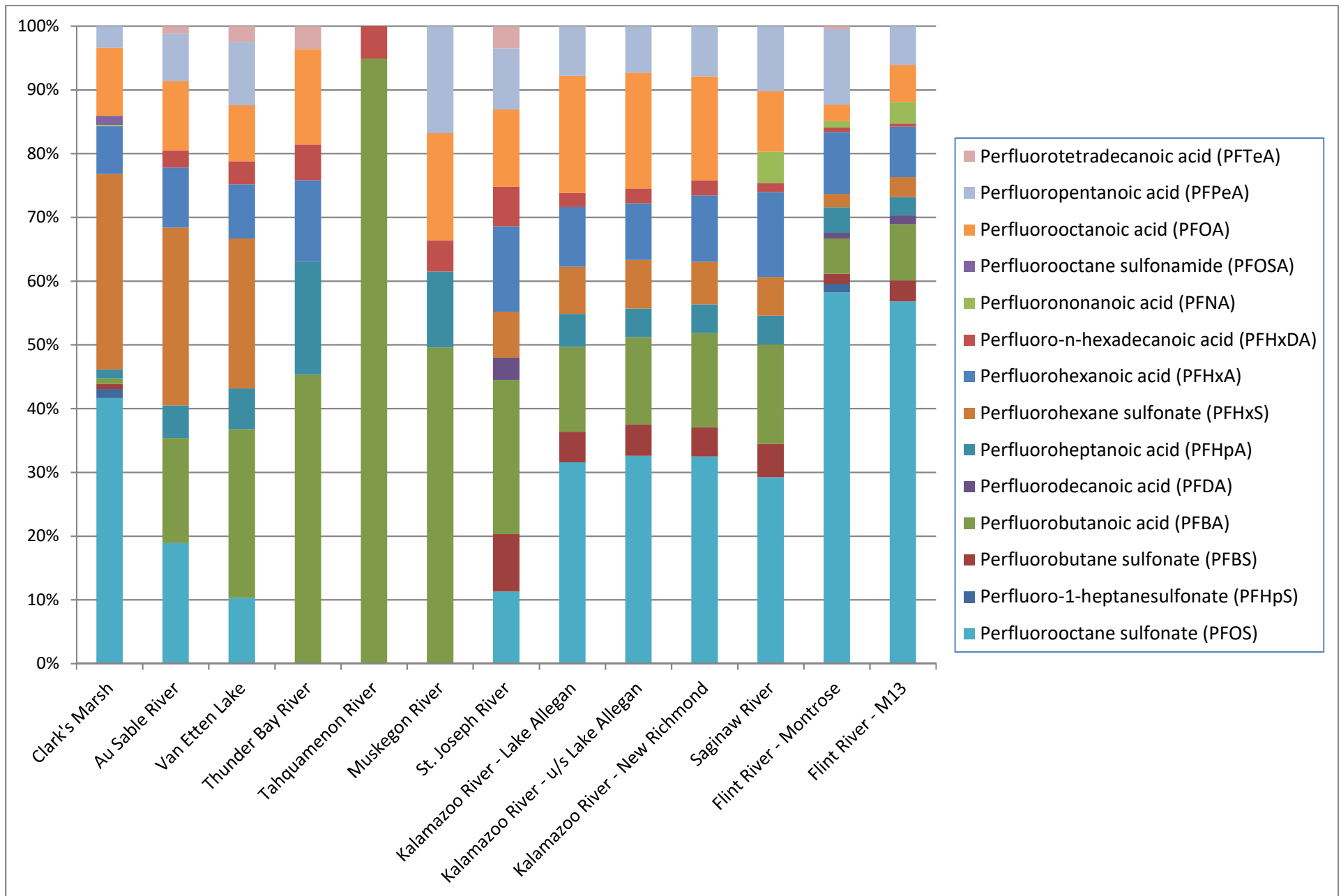


Figure 2. Individual PFCs in water samples as a percentage of the sum of all PFCs measured.

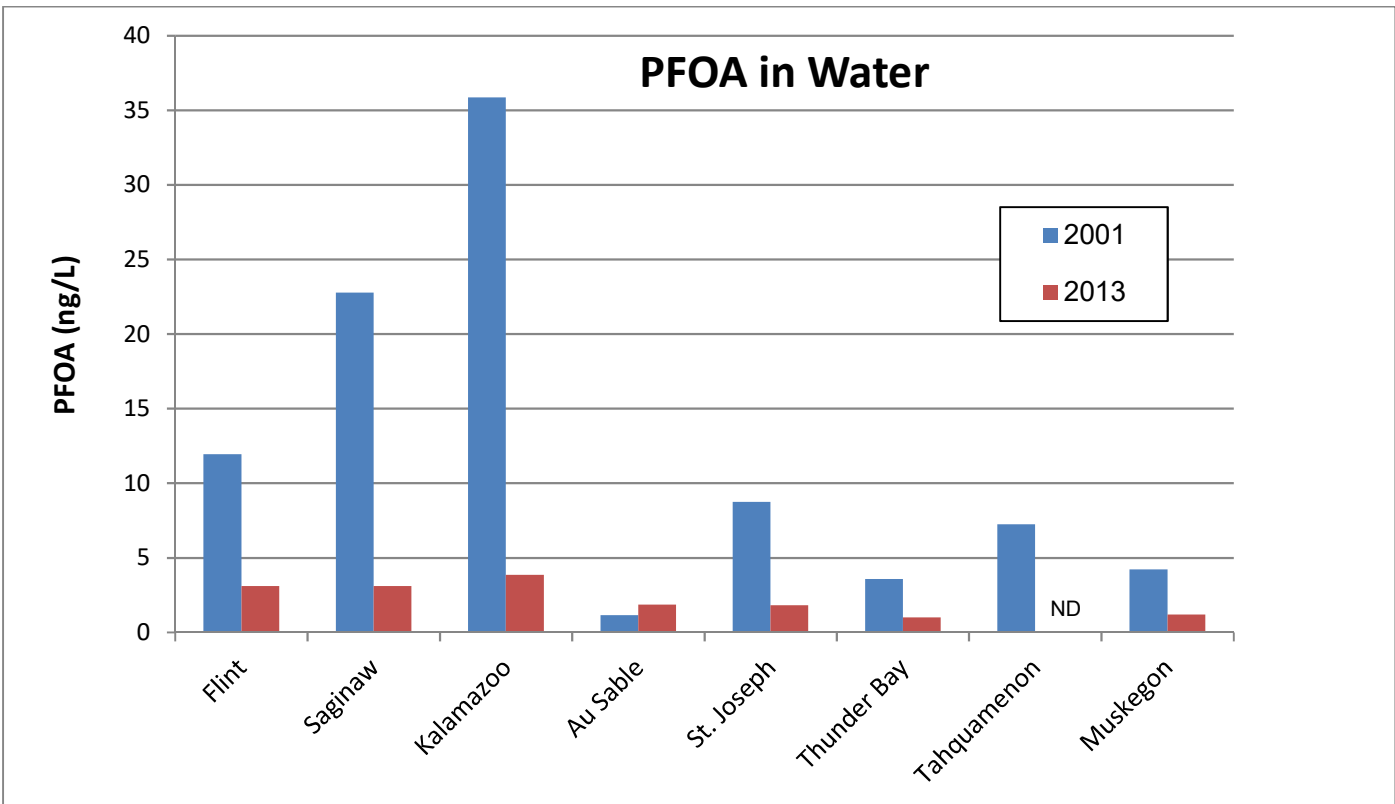
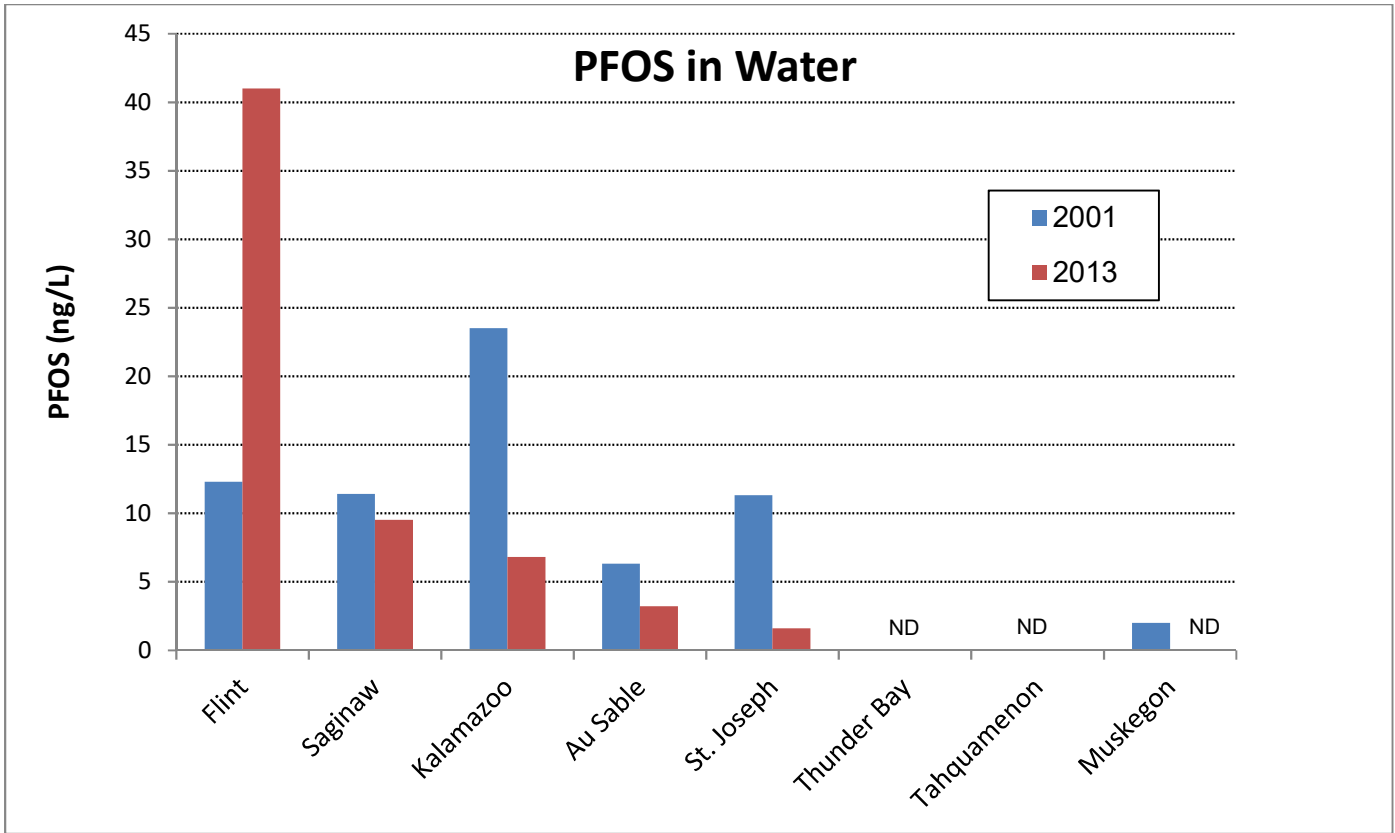


Figure 3. A comparison of PFOS and PFOA concentrations in water samples collected at selected sites in 2001 and 2013. ND indicates concentrations are below the detection limit.

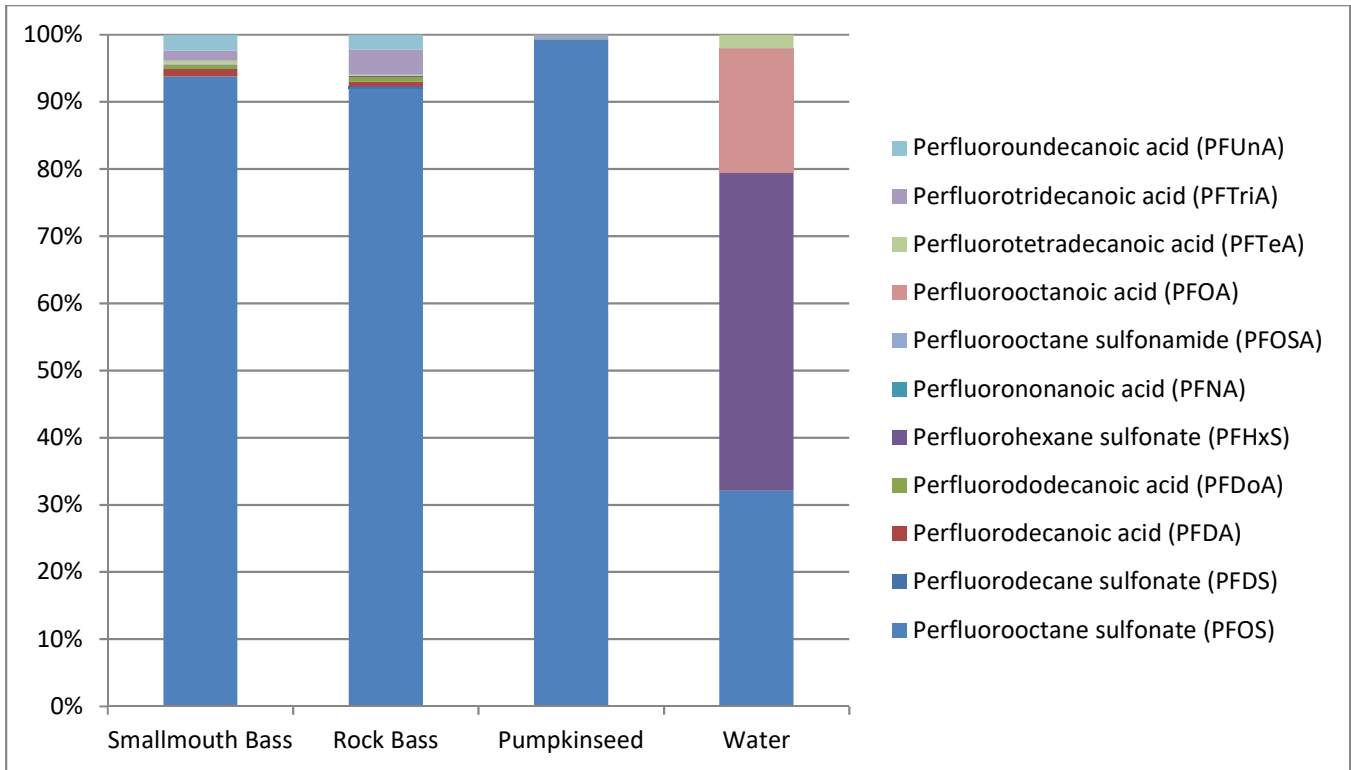


Figure 4. Individual PFCs in water and fish samples collected from the Au Sable River between 2011 and 2013, as a percentage of the 11 PFCs measured in both fish and water.

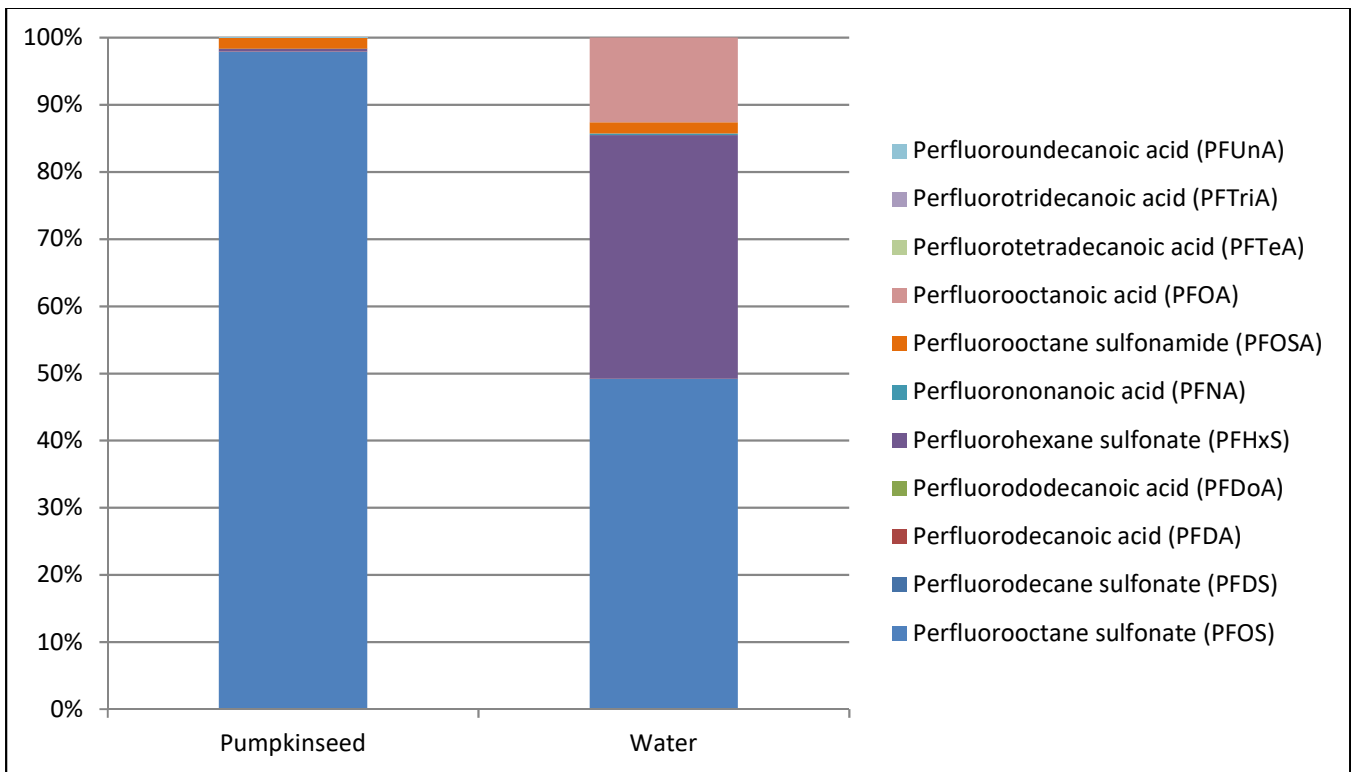


Figure 5. Individual PFCs in water and fish samples collected from the Clark's Marsh in 2011, as a percentage of the 11 PFCs measured in both fish and water.

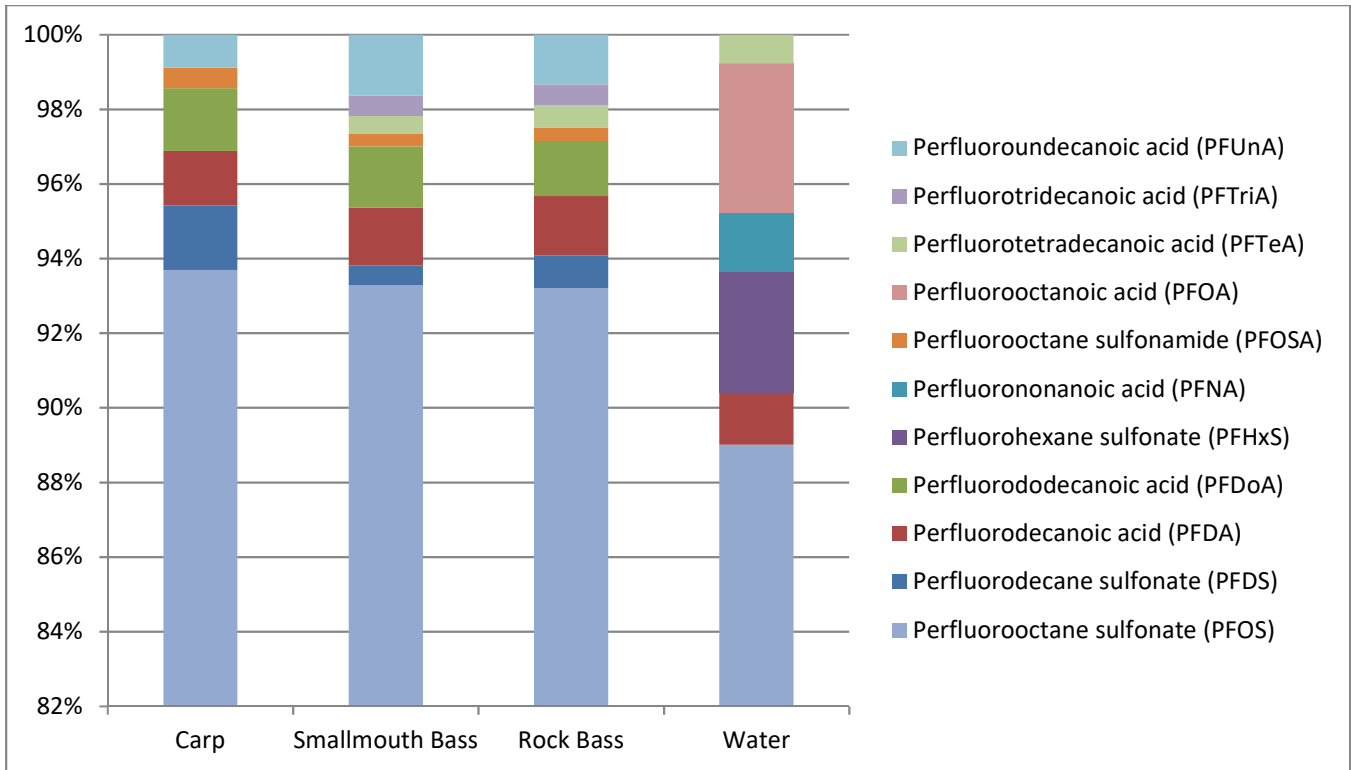


Figure 6. Individual PFCs in water and fish samples collected from the Flint River in 2013, as a percentage of the 11 PFCs measured in both fish and water.

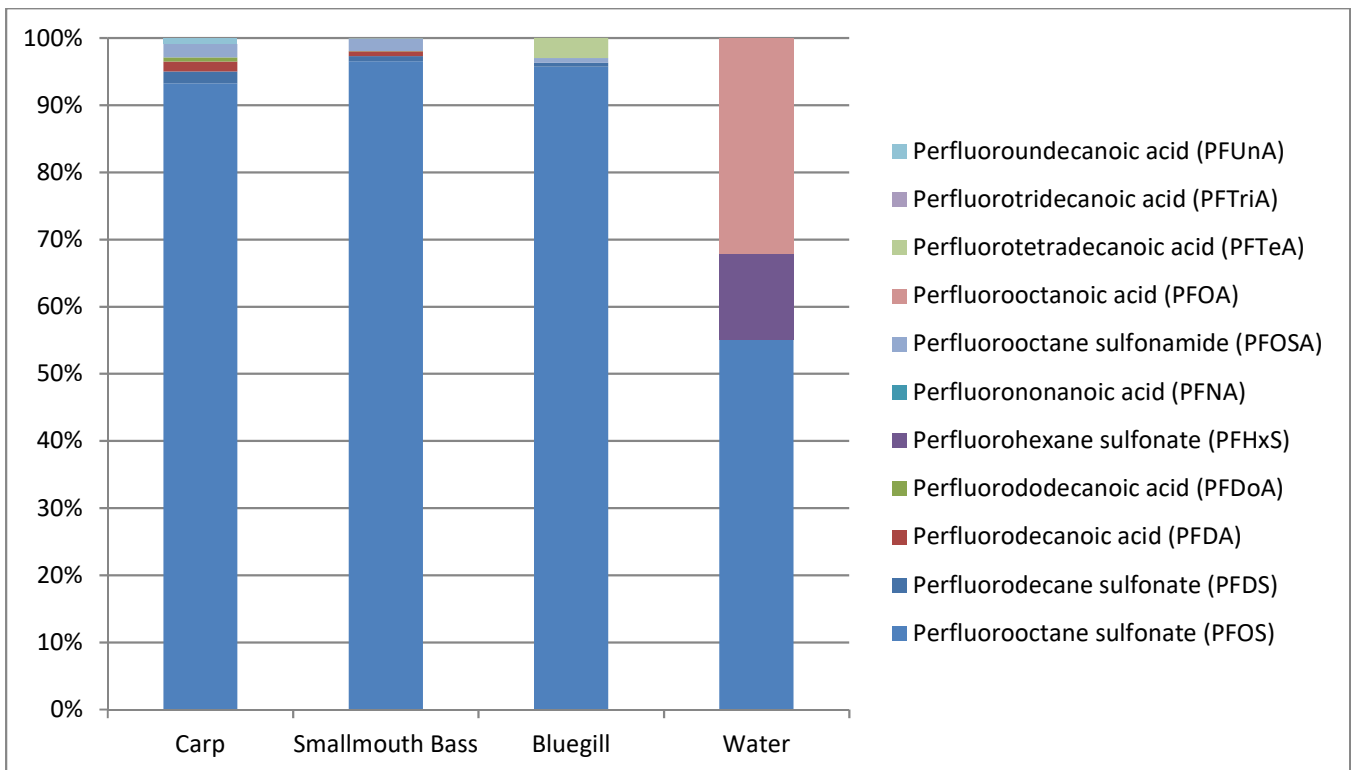


Figure 7. Individual PFCs in water and fish samples collected from the Kalamazoo River at Lake Allegan in 2013, as a percentage of the 11 PFCs measured in both fish and water.

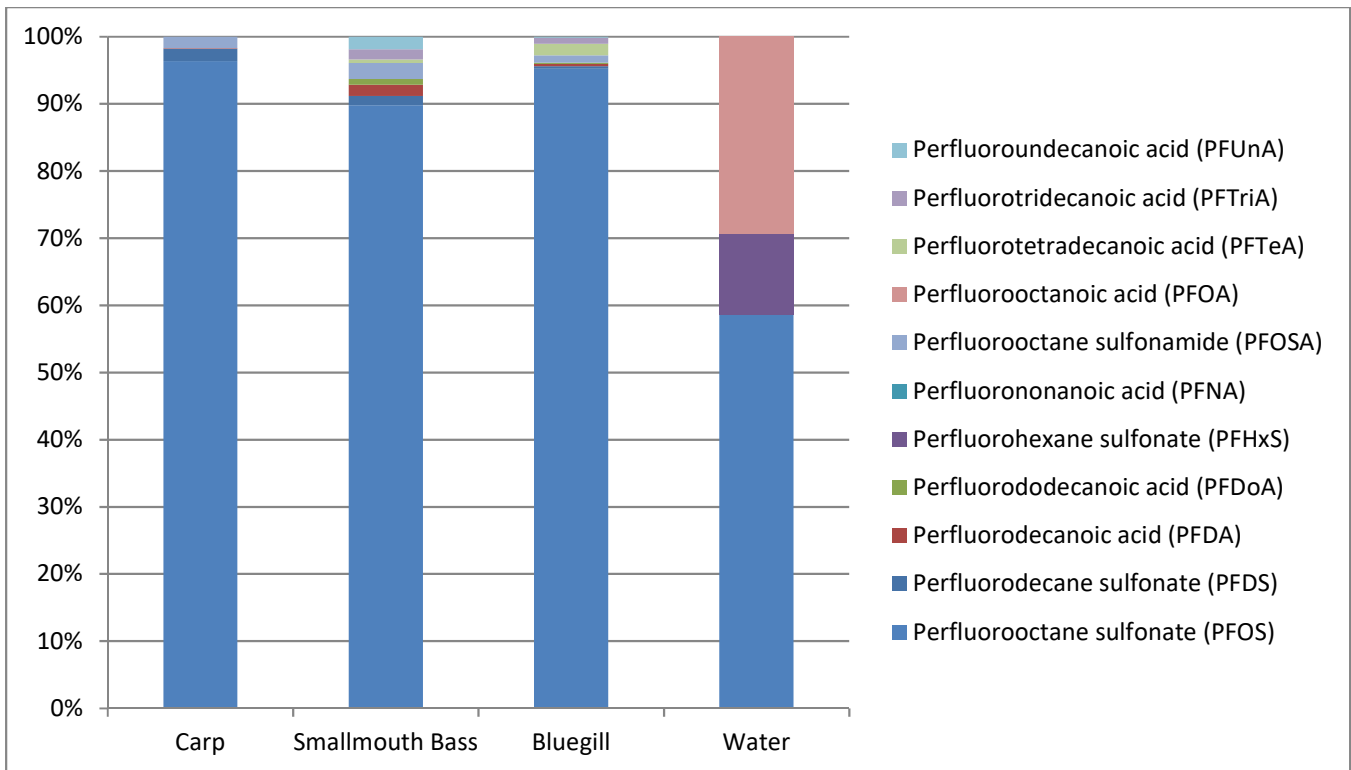


Figure 8. Individual PFCs in water and fish samples collected from the Kalamazoo River at New Richmond in 2013, as a percentage of the 11 PFCs measured in both fish and water.

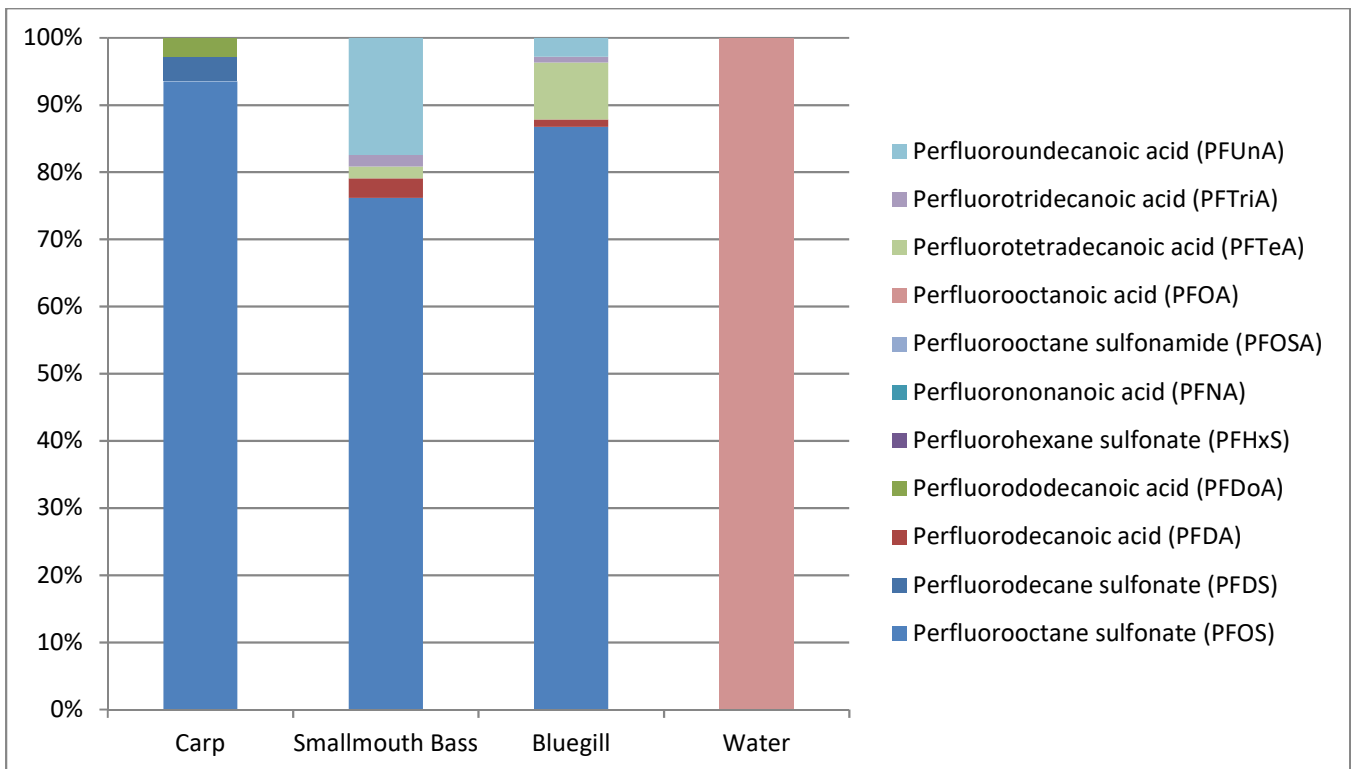


Figure 9. Individual PFCs in water and fish samples collected from the Muskegon River in 2013, as a percentage of the 11 PFCs measured in both fish and water.

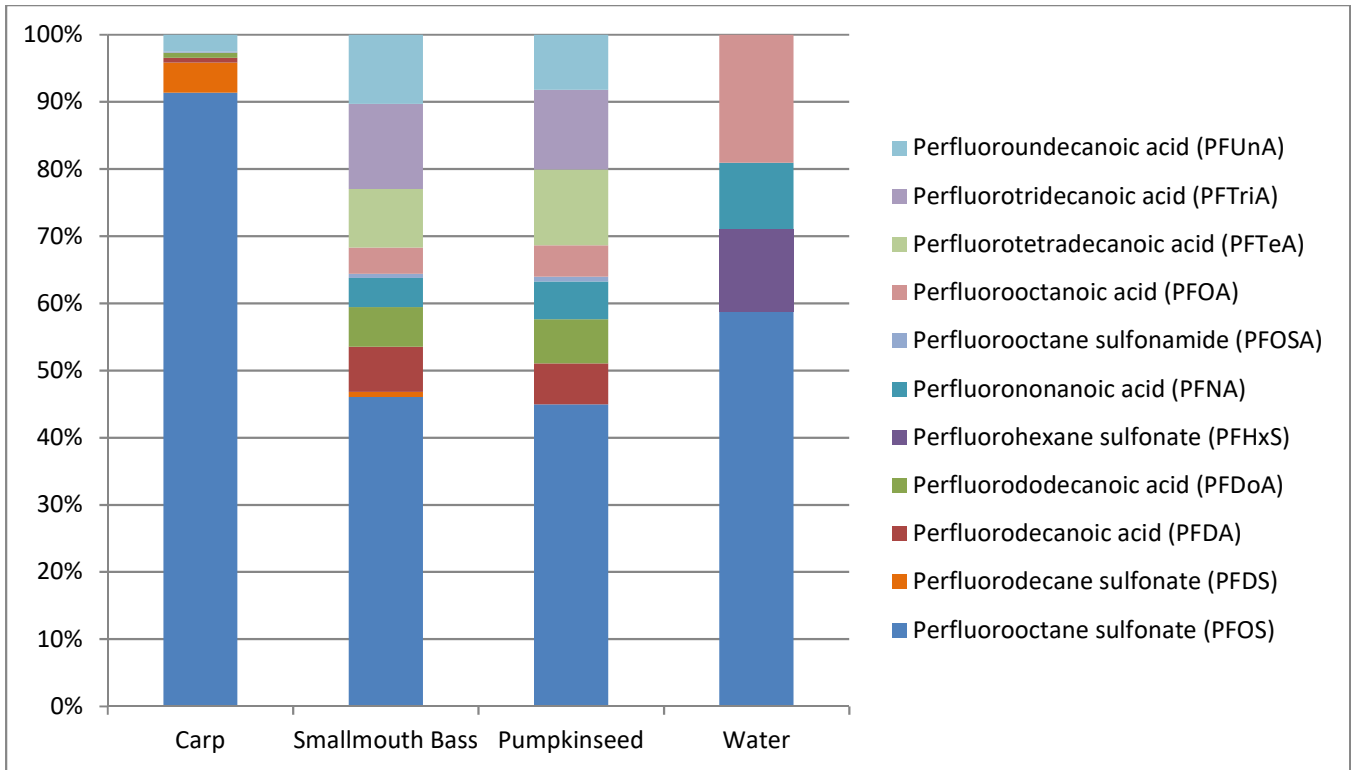


Figure 10. Individual PFCs in water and fish samples collected from the Saginaw River in 2013, as a percentage of the 11 PFCs measured in both fish and water.

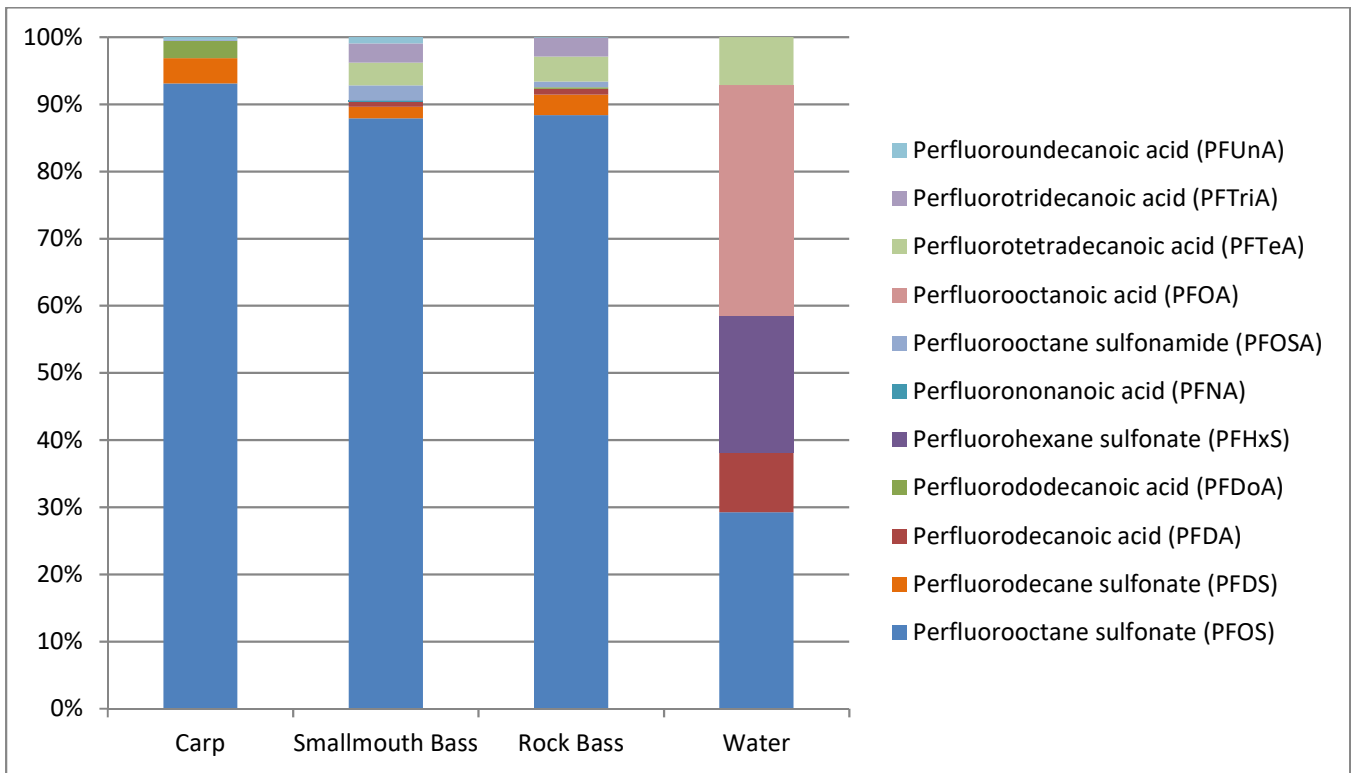


Figure 11. Individual PFCs in water and fish samples collected from the St. Joseph River in 2013, as a percentage of the 11 PFCs measured in both fish and water.

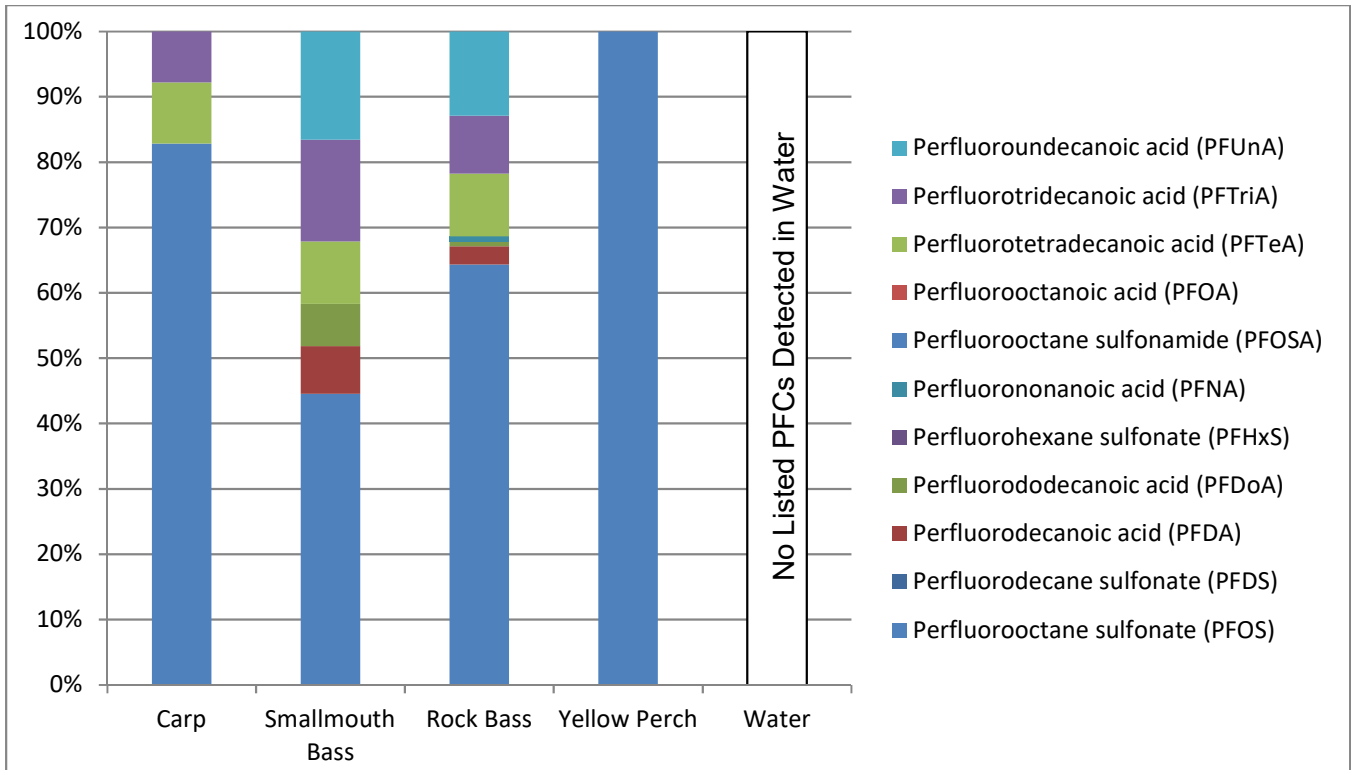


Figure 12. Individual PFCs in water and fish samples collected from the Tahquamenon River in 2013 and 2014, as a percentage of the 11 PFCs measured in both fish and water.

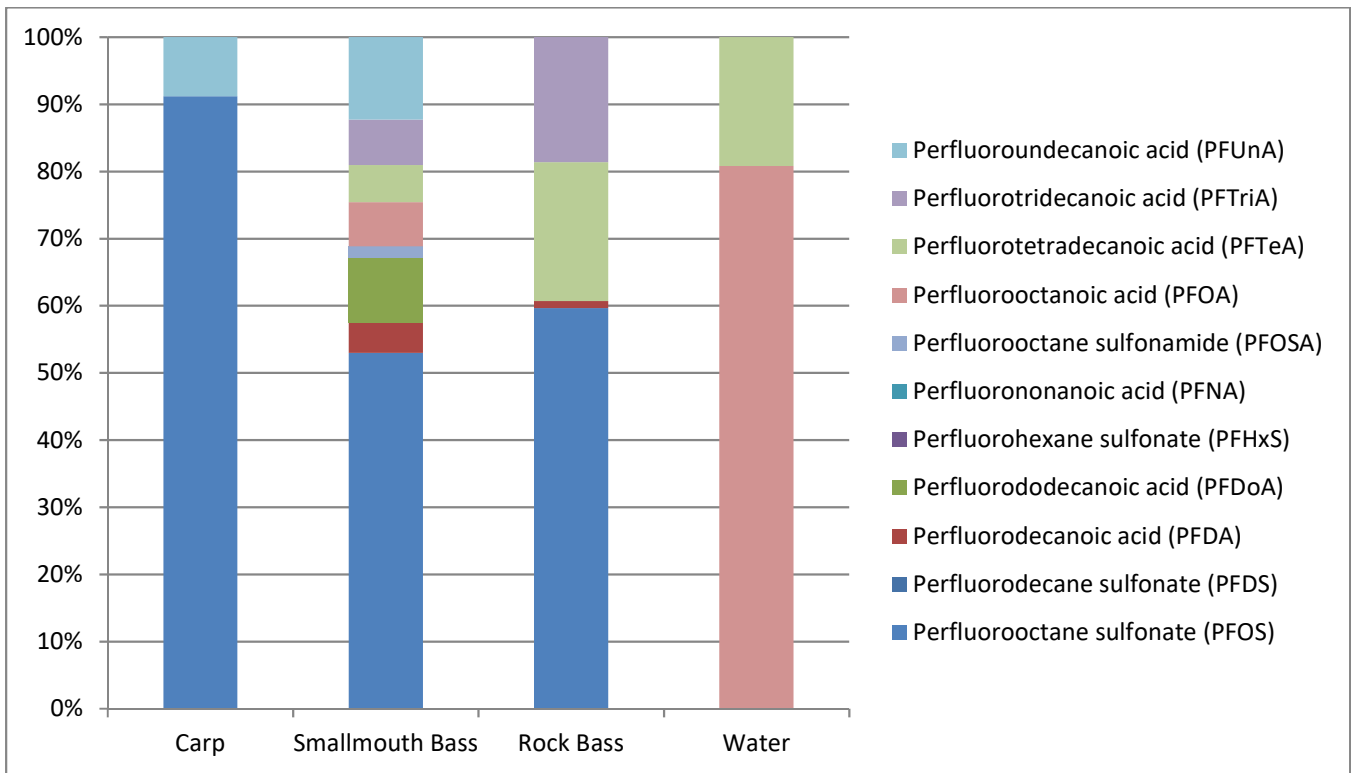


Figure 13. Individual PFCs in water and fish samples collected from the Thunder Bay River at Lake Besser in 2013, as a percentage of the 11 PFCs measured in both fish and water.

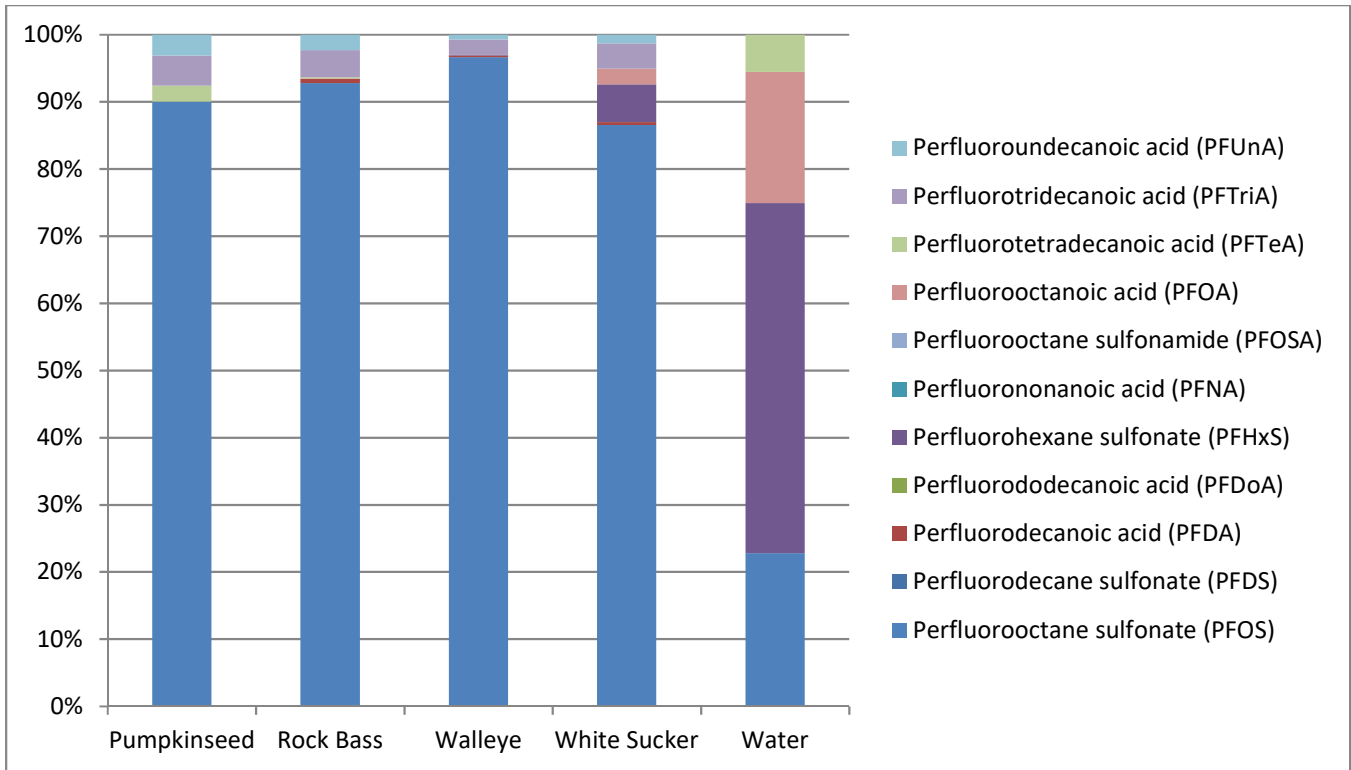


Figure 14. Individual PFCs in water and fish samples collected from Van Etten Lake between 2010 and 2013, as a percentage of the 11 PFCs measured in both fish and water.

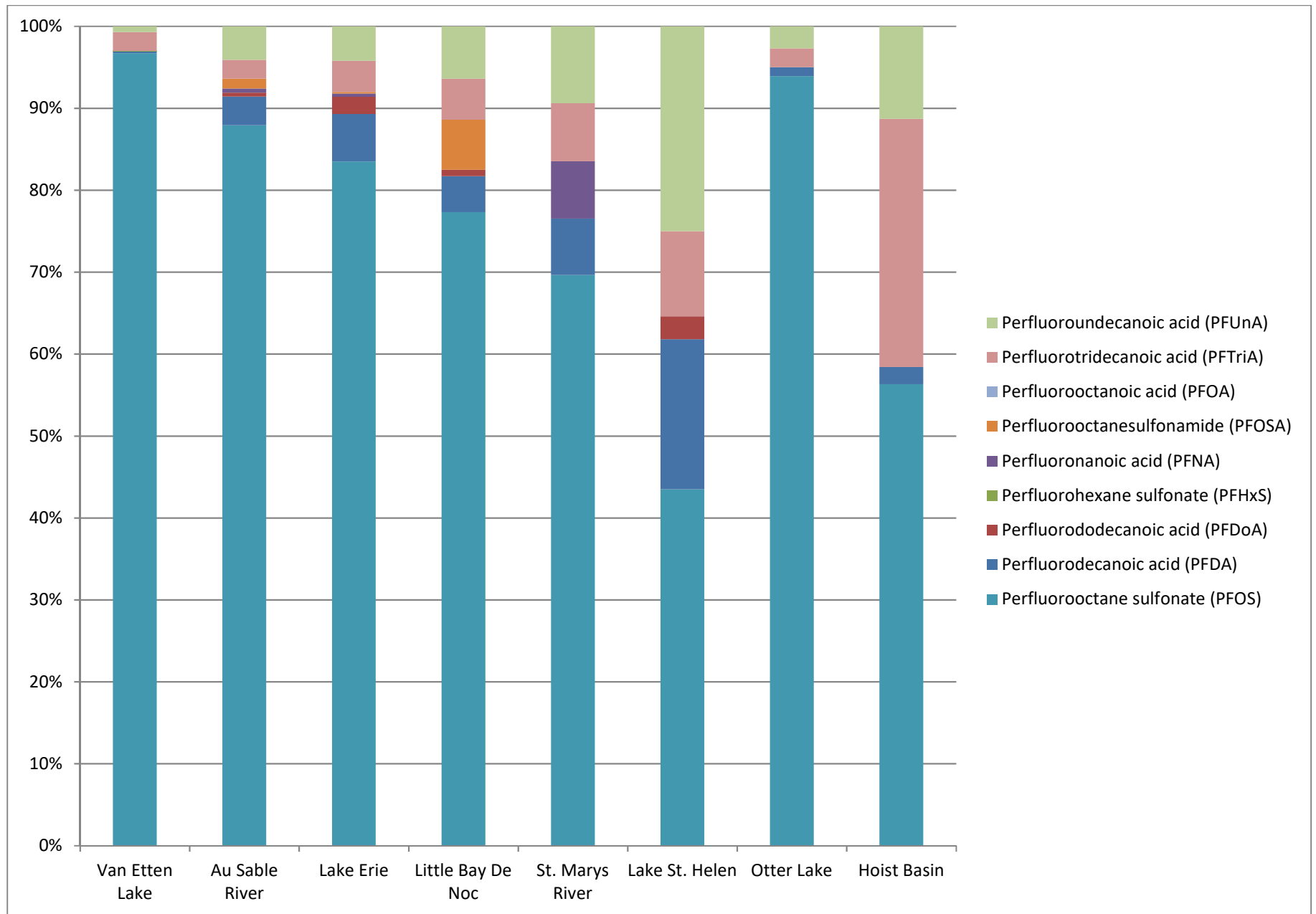


Figure 15. Individual PFCs in walleye fillet samples collected from Michigan waters in 2010-2014, as a percentage of all PFCs measured.

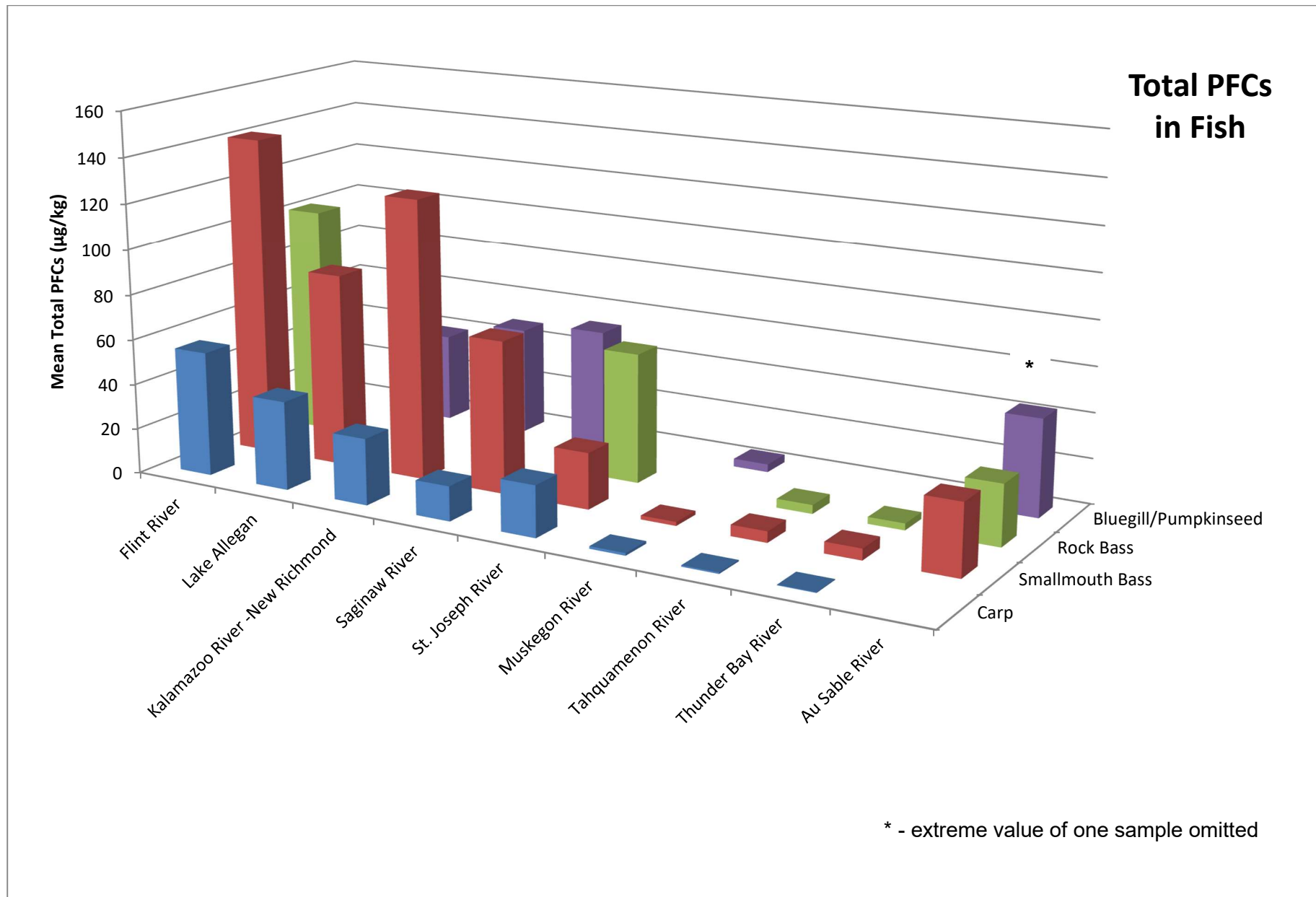


Figure 16. Mean concentrations of total PFCs in fillets of fish from 9 sites sampled in 2013.

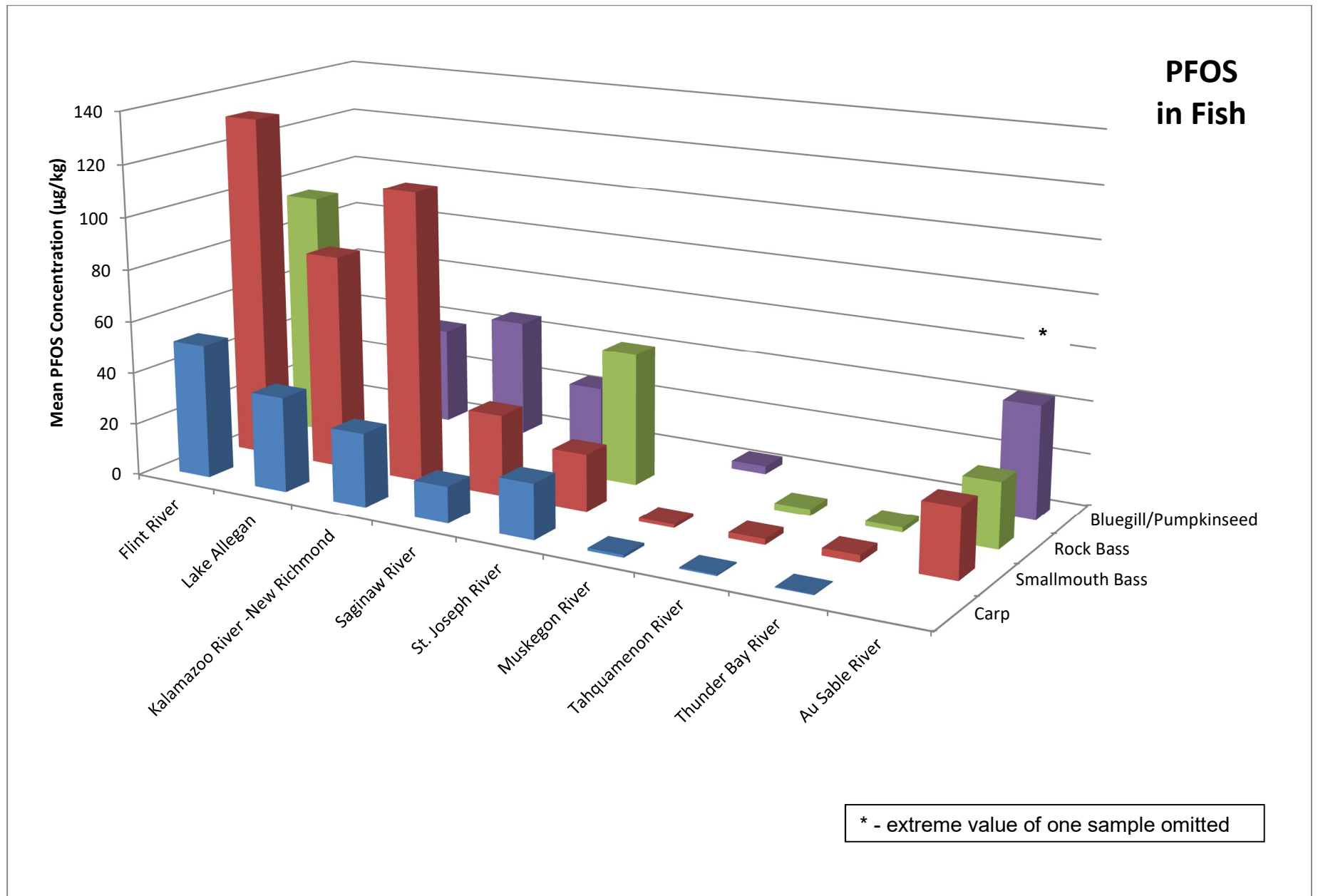


Figure 17. Mean concentrations of PFOS in fillets of fish from 9 sites sampled in 2013.

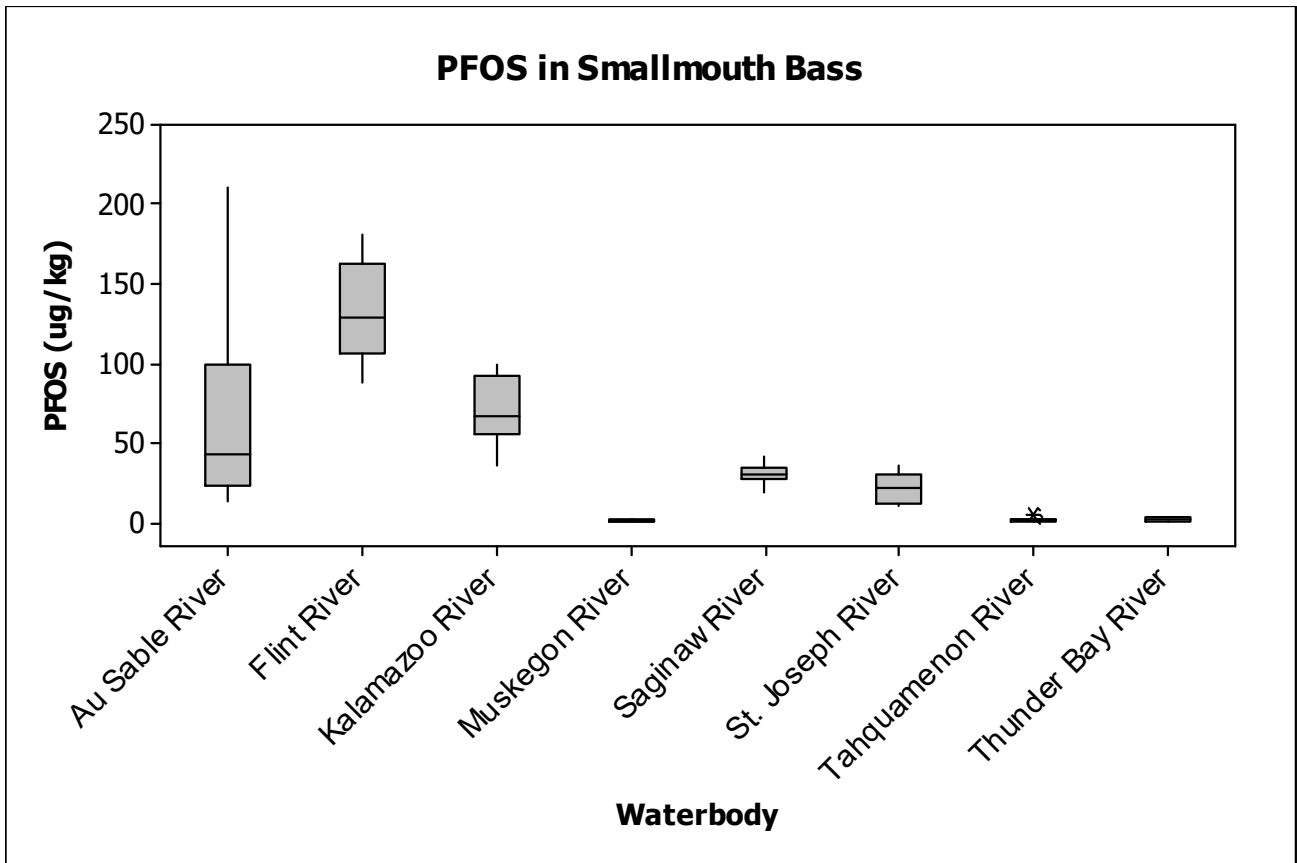


Figure 18. Boxplot of PFOS concentrations in fillets of smallmouth bass collected in 2013. (* denotes a value beyond 1.5 times the inter-quartile range).

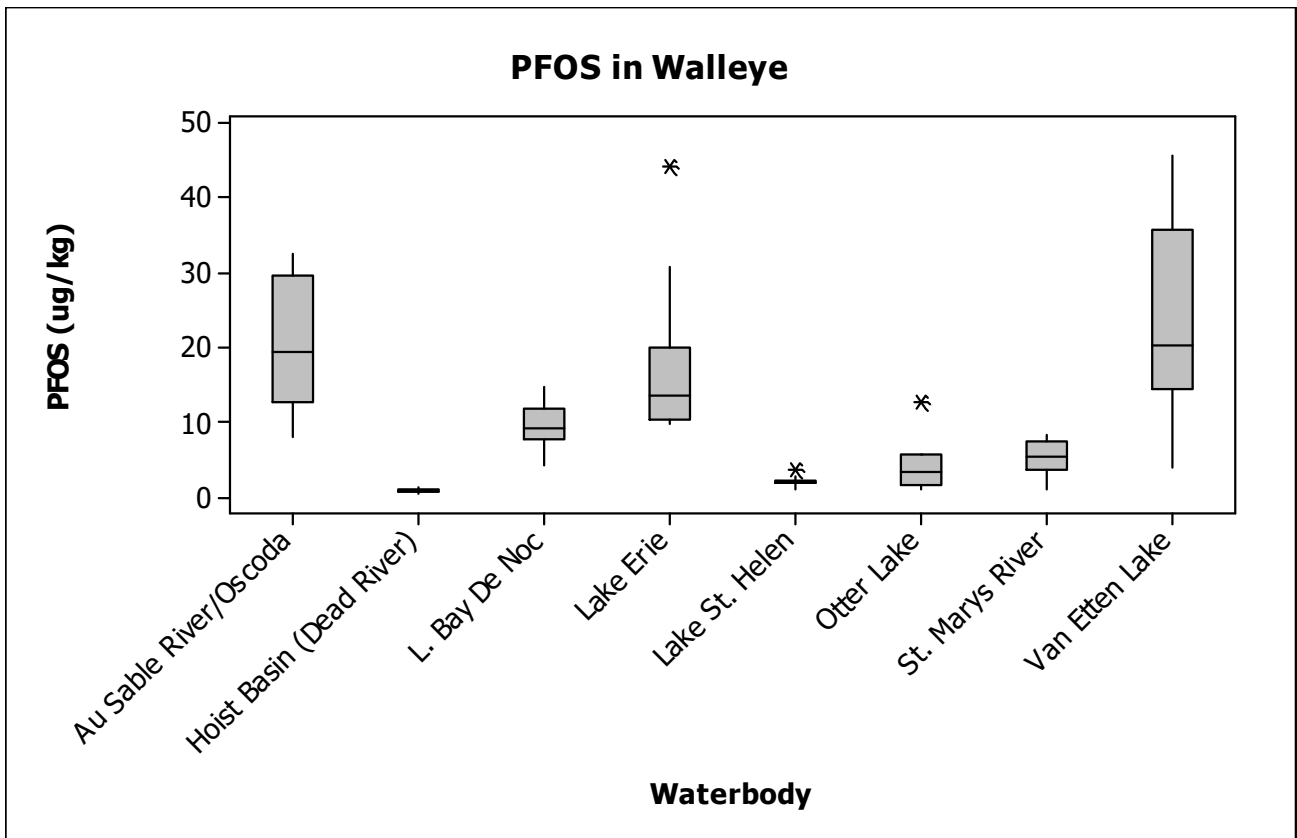


Figure 19. Boxplot of PFOS concentrations in fillets of walleye collected at selected sites in Michigan waters. (* denotes a value beyond 1.5 times the inter-quartile range).

**MICHIGAN DEPARTMENT OF ENVIRONMENTAL QUALITY
WATER RESOURCES DIVISION
HUMAN NONCANCER VALUE WORKSHEET**

Chemical Name: Perfluorooctane sulfonate (PFOS)
 Developed By: A. Babcock
 Reviewed By: D. Bush

CAS No.: 1763-23-1
 Literature Review Date: 7/29/2013
 Verification Date: 3/20/2014

Key Study: Seacat et al. (2002) dosed male and female cynomolgus monkeys with PFOS potassium salt capsules daily for 182 days. The NOAEL was 0.03 mg/kg/d. Effects at the higher doses (0.15 mg/kg/d and 0.75 mg/kg/d) included decreased body weight, increased liver weight, and decreased levels of serum total cholesterol, triiodothyronine (T₃), and estradiol. Additionally, in the highest dose group, one male died and one male was subject to early moribund sacrifice during the treatment phase of the study.

Physiologically-based pharmacokinetic modeling predicts the area under the curve (AUC) for the NOAEL to be 22,100 mg/L*h (see justification for details). The AUC is converted to a human equivalent dose (HED) by multiplying by the human clearance of PFOS (calculated to be 0.000081 L/kg/d). Thus, the HED is 0.00041 mg/kg/d.

A total uncertainty factor (UF) of 30 is applied to the HED to account for interspecies toxicodynamics (3x) and intraspecies variability (10x). Despite the subchronic study duration, the UF for that extrapolation is set to 1x because the PBPK model accounts for that uncertainty in the determination of the steady-state serum concentration.

$$\text{ADE} = \frac{0.00041 \text{ mg/kg/d}}{30}$$

Where UF = 10x for intraspecies extrapolation and 3x for interspecies extrapolation.

$$\text{ADE} = 0.000013667 \text{ mg/kg/d}$$

drinking water

$$\text{HNV} = \frac{0.000013667 \text{ mg/kg/d} \times (70 \text{ kg}) \times (0.8) \times 1000}{(2 \text{ L/d}) + (0.0036 \text{ kg/d} \times 2329 \text{ L/kg}) + (0.0114 \text{ kg/d} \times 5047 \text{ L/kg})} = 0.01127 \text{ ug/L}$$

HNV for drinking water = 0.011 ug/L

non-drinking water

$$\text{HNV} = \frac{0.000013667 \text{ mg/kg/d} \times (70 \text{ kg}) \times (0.8) \times 1000}{(0.01 \text{ L/d}) + (0.0036 \text{ kg/d} \times 2329 \text{ L/kg}) + (0.0114 \text{ kg/d} \times 5047 \text{ L/kg})} = 0.0116 \text{ ug/L}$$

HNV for non-drinking water = 0.012 ug/L

**MICHIGAN DEPARTMENT OF ENVIRONMENTAL QUALITY
WATER RESOURCES DIVISION
HUMAN NONCANCER VALUE WORKSHEET**

Chemical Name: Perfluorooctanoic acid (PFOA)
 Developed By: A. Babcock
 Reviewed By: D. Bush

CAS No.: 335-67-1
 Literature Review Date: 9/1/2010
 Verification Date: 5/5/2011

Key Study: Butenhoff et al. (2002) administered ammonium perfluorooctanoate to male cynomolgus monkeys via oral capsule daily for 26 weeks. Treatment-related effects were noted in all treatment groups; the LOAEL for the study was 3 mg/kg/d, based on increased absolute liver weight. The study LOAEL is adjusted to account for the different body burdens expected in humans versus monkeys owing to the orders-of-magnitude longer half-life of PFOA in humans than in monkeys. The resultant human LOAEL is 0.046 mg/kg/d and is divided by a total uncertainty factor of 3000. Additional details are provided in the toxicological assessment.

$$\text{ADE} = \frac{0.046 \text{ mg/kg/d}}{3,000}$$

Where UF = 10x each for subchronic-to-chronic, LOAEL-to-NOAEL and intraspecies extrapolation, and 3x for interspecies extrapolation.

$$\text{ADE} = 0.0000153 \text{ mg/kg/d}$$

drinking water:

$$\text{HNV} = \frac{(0.0000153 \text{ mg/kg/d}) \times (70 \text{ kg}) \times (0.8)}{(2 \text{ L/d}) + (0.0036 \text{ kg/d}) \times (4 \text{ L/kg}) + (0.0114 \text{ kg/d}) \times (4 \text{ L/kg})} = 0.4168 \text{ ug/L}$$

$$\text{HNV for drinking water} = 0.42 \text{ ug/L}$$

non-drinking water:

$$\text{HNV} = \frac{(0.0000153 \text{ mg/kg/d}) \times (70 \text{ kg}) \times (0.8)}{(0.01 \text{ L/d}) + (0.0036 \text{ kg/d}) \times (4 \text{ L/kg}) + (0.0114 \text{ kg/d}) \times (4 \text{ L/kg})} = 12.2667 \text{ ug/L}$$

$$\text{HNV for non-drinking water} = 12 \text{ ug/L}$$

**MICHIGAN DEPARTMENT OF ENVIRONMENTAL QUALITY
WATER RESOURCES DIVISION
WILDLIFE VALUE WORKSHEET**

Chemical Name: Perfluorooctane sulfonate CAS No. 1763-23-1
 Developed By: D. Bush
 Reviewed By: D. Bush Verification Date: 4/9/15

Key Study:

Mallard ducks and bobwhite quail were exposed to 0, 10, 50, or 150 mg PFOS/kg diet for 21 weeks (Newsted et al., 2007). Due to an increase in mortality, the ducks and quail exposed to 50 and 150 mg/kg were terminated by week 7. The LOAEL for quail (decreased survivorship of 14-day-old offspring) was 10 mg/kg. A slightly greater incidence of small testes occurred in ducks (and quail) in the 10 mg/kg exposure group. However, since no effects were observed on duck reproductive performance, this dose was considered a NOAEL for population-level effects in ducks. The LOAELs for quail and ducks were reported as 0.77 and 1.49 mg/kg bwt/d, respectively (Newsted et al., 2007). The Test Dose (TD) of 0.77 mg/kg bwt/d was used to derive the avian Wildlife Value (WV).

$$WV = [(TD/Uncertainty\ Factor)(Weight)] / [(Water\ Consumption) + \sum(F_{TLi} \times BAF_{TLi})]$$

Uncertainty Factor = 30, consisting of the following individual uncertainty factors:

- A 10-fold uncertainty factor was used for interspecies extrapolation because toxicity data were only available for two species of birds.
- A 3-fold uncertainty factor was used to extrapolate from a LOAEL to a NOAEL. A factor of 10 was not used because a pilot study (Newsted et al., 2005) found no adverse effects at 6.2 mg/kg, suggesting that 10 mg/kg is near the threshold.
- No additional uncertainty factor was used to extrapolate from a 21-week study to a chronic study because the key study looked at a sensitive endpoint.

$$WV_{\text{eagles}} = [(0.77 \text{ mg/kg bwt/d} \div 30)(4.6 \text{ kg})] \div [(0.160 \text{ L/d}) + [(0.371 \text{ kg/d} \times 2,367 \text{ L/kg}) + (0.0928 \text{ kg/d} \times 5,129 \text{ L/kg}) + (0.0283 \text{ kg/d} \times 5,129 \times 9.0^*)]]$$

$$WV_{\text{eagles}} = 0.000044335 \text{ mg/L} = 0.044 \text{ ug/L}$$

*The ratio of PFOS levels in glaucous gull livers to whole arctic cod was found to be 9.0 (Tomy et al., 2004). This value is supported by the value of 8.9 found when comparing the concentration of PFOS in the livers of mergansers to the concentration of PFOS in the livers of fish (smallmouth and largemouth bass) from the Niagara region of New York (Sinclair et al., 2006).

$$WV_{\text{gulls}} = [(0.77 \text{ mg/kg bwt/d} \div 30)(1.1 \text{ kg})] \div [(0.063 \text{ L/d}) + [(0.192 \text{ kg/d} \times 2,367 \text{ L/kg}) + (0.0480 \text{ kg/d} \times 5,129)]]$$

$$WV_{\text{gulls}} = 0.000040273 \text{ mg/L} = 0.040 \text{ ug/L}$$

$$WV_{\text{kingfisher}} = [(0.77 \text{ mg/kg bwt/d} \div 30)(0.15 \text{ kg})] \div [(0.017 \text{ L/d}) + [(0.0672 \text{ kg/d} \times 2,367 \text{ L/kg})]]$$

$$WV_{\text{kingfisher}} = 0.000024202 \text{ mg/L} = 0.024 \text{ ug/L}$$

$$WV_{\text{avian}} = (0.044 \times 0.040 \times 0.024)^{1/3} = 0.0348 \text{ ug/L} = 0.035 \text{ ug/L}$$

References:

Newsted, J.L., K.K. Coady, S.A. Beach, J.L. Butenhoff, S. Gallagher, and J.P. Giesy. 2007. Effects of perfluorooctane sulfonate on mallard and northern bobwhite quail exposed chronically via the diet. *Environ. Toxicol. Pharmacol.* 23:1-9.

Newsted, J.L., P.D. Jones, K. Coady, and J.P. Giesy. 2005. Avian toxicity reference values for perfluorooctane sulfonate. *Environ. Sci. Technol.* 39:9357-9362.

Sinclair, E., D.T. Mayack, K. Roblee, N. Yamashita, and K. Kannan. 2006. Occurrence of perfluoroalkyl surfactants in water, fish, and birds from New York state. *Arch. Environ. Contam. Toxicol.* 50:398-410.

Tomy, G.T., W. Budakowski, T. Halldorson, P.A. Helm, G.A. Stern, K. Friesen, K. Pepper, S.A. Tittlemier, and A.T. Fisk. 2004. Fluorinated organic compounds in the eastern arctic marine food web. *Environ. Sci. Technol.* 38:6475-6481.

**MICHIGAN DEPARTMENT OF ENVIRONMENTAL QUALITY
WATER RESOURCES DIVISION
WILDLIFE VALUE WORKSHEET**

Chemical Name: Perfluorooctane sulfonate CAS No. 1763-23-1
 Developed By: D. Bush
 Reviewed By: [Signature] Verification Date: 4/9/05

Key Study:

Since no studies were found that exposed mammalian wildlife to PFOS, it was necessary to base the mammalian Wildlife Value (WV) on toxicity studies conducted on laboratory animals. A 2-generation rat study (Luebker et al., 2005) was considered the best study for WV derivation. In this study, male and female rats were dosed via oral gavage at dose levels of 0, 0.1, 0.4, 1.6, and 3.2 mg/kg/d for 6 weeks prior to mating, during mating, and for females through gestation and lactation, across two generations. Substantial neonatal toxicity occurred in the 1.6 and 3.2 mg/kg/d dose groups. The slight delay in eye opening in F₁ pups and the transient reduction in body weights in F₂ pups in the 0.4 mg/kg treatment group were not considered toxicologically significant by the researchers. The weight decrease that occurred in the F₀ generation adults in the 0.4 mg/kg treatment group was also not considered significant because there was a decrease in food consumption at this dose and the effect did not occur in the F₁ generation adults. The dose of 0.4 mg/kg/d was considered a NOAEL for effects on wildlife populations and was used to derive the mammalian WV.

$$WV = [(TD/Uncertainty\ Factor)(Weight)] / [(Water\ Consumption) + \Sigma(F_{TLi} \times BAF_{TLi})]$$

The key study was of chronic duration and identified a NOAEL, thus the only uncertainty factor necessary is that for interspecies extrapolation. A factor of 10 was used for this UF since no mammalian wildlife data were available.

$$WV_{mink} = [(0.4\ mg/kg\ bwt/d \div 10)(0.8\ kg)] \div [(0.081\ L/d) + [(0.159\ kg/d \times 2,367\ L/kg) + (0 \times 5,129)]]$$

$$WV_{mink} = 0.000085008\ mg/L = 0.085\ ug/L$$

$$WV_{otter} = [(0.4\ mg/kg\ bwt/d \div 10)(7.4\ kg)] \div [(0.60\ L/d) + [(0.976 \times 2,367\ L/kg) + 0.244\ kg/d \times 5,129]]$$

$$WV_{otter} = 0.000083093\ mg/L = 0.083\ ug/L$$

$$WV_{mammalian} = (0.085 \times 0.083)^{1/2} = 0.084\ ug/L$$

References:

Luebker, D.J., M.T. Case, R.G. York, J.A. Moore, K.J. Hansen, and J.L. Butenhoff. 2005. Two-generation reproduction and cross-foster studies of perfluorooctanesulfonate (PFOS) in rats. Toxicol. 215:126-148.